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Charlton Lane Eco Park



SUEZ Recycling and Recovery Surrey Ltd

Dioxin Pathway Risk Assessment

Document approval

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1 Introduction

Fichtner Consulting Engineers Ltd (“Fichtner”) has been engaged by SUEZ Recycling and Recovery Surrey Ltd to undertake a Dioxin Pathway Intake Assessment to support the application for a variation to the Environmental Permit (EP) for the Charlton Lane Eco Park (herein referred to as the Facility). Full details of the proposed changes being applied for can be found in the Supporting Information document.

When considering the impact on human health, the predicted atmospheric concentrations have been compared to the Air Quality Assessment Levels (AQALs) for the protection of human health within the Dispersion Modelling Assessment to accompany the EP application. It is noted that for dioxins the AQAL is a Tolerable Daily Intake (TDI) which considered the combination of the intake from inhalation and ingestion. As such it is not possible to demonstrate compliance with the assessment level with just reference to the air concentration. As such, this assessment has been undertaken to assess the pathway intake of these pollutants and impacts compared to the TDI. The Facility includes gasification plant, and an anaerobic digestion plant with associated biogas engines and flare. Dioxins are only released from the gasification plant; therefore, this assessment has only considered the impact of emissions from the gasification plant.

2 Literature review

The Health Protection Agency (HPA), whose role was taken over by Public Health England (PHE) and more recently the UK Health Security Agency (HAS), published a note RCE-13 “The Impact on Health of Emissions to Air from Municipal Waste Incinerators”, in 2009¹. The summary states:

“While it is not possible to rule out adverse health effects from modern, well-regulated municipal waste incinerators with complete certainty, any potential damage to the health of those living close-by is likely to be very small, if detectable”

PHE commissioned further research in 2012, while continuing to state that the conclusions of RCE-13 remain applicable. These studies were commissioned from the Small Area Health Statistics Unit, which is based at Imperial College London and Kings College London. The methodology and results of the studies have been published in a series of papers in scientific journals. The three most recent papers, known as Ghosh et al (2018)², Freni-Sterrantino et al (2019)³ and Parkes et al (2019)⁴, are the most relevant.

These studies considered whether living near a municipal waste incinerator (MWI) is linked with adverse reproductive and infant health outcomes. These outcomes were studied as they are considered more sensitive to the accumulation of pollutants in the environment than other potential markers such as lifetime cancer rates.

Ghosh et al (2018) concluded that:

“This large national study found no evidence for increased risk of a range of birth outcomes, including birth weight, preterm delivery and infant mortality, in relation to either MWI emissions or living near an MWI operating to the current EU waste incinerator regulations in Great Britain.”

Freni-Sterrantino et al (2019) concluded that:

“we did not find an association between the opening of a new MWI and changes in infant mortality trends or sex ratio at birth for 10 and 4 km buffers, using distance as proxy of exposure, after taking into account temporal trends in comparator areas and potential confounding factors.”

The objective of Parkes et al (2019) was as follows: *“To conduct a national investigation into the risk of congenital anomalies in babies born to mothers living within 10 km of an MWI associated with: i) modelled concentrations of PM₁₀ as a proxy for MWI emissions more generally and; ii) proximity of residential postcode to nearest MWI, in areas in England and Scotland that are covered by a congenital anomaly register.”* Under objective (i), which related congenital anomalies to modelled concentrations and so would be considered the more representative approach, the study

¹ <https://www.gov.uk/government/publications/municipal-waste-incinerators-emissions-impact-on-health>

² Ghosh RE, Freni Sterrantino A, Douglas P, Parkes B, Fecht D, de Hoogh K, Fuller G, Gulliver J, Font A, Smith RB, Blangiardo M, Elliott P, Toledano MB, Hansell AL. (2018) Fetal growth, stillbirth, infant mortality and other birth outcomes near UK municipal waste incinerators; retrospective population based cohort and case-control study. Environment International.

³ Freni-Sterrantino, A; Ghosh, RE; Fecht, D; Toledano, MB; Elliott, P; Hansell, AL; Blangiardo, M. (2019) Bayesian spatial modelling for quasi-experimental designs: An interrupted time series study of the opening of Municipal Waste Incinerators in relation to infant mortality and sex ratio. Environment International. 128 106-115

⁴ Parkes B, Hansell A.L., Ghosh R.E, Douglas P., Fecht D., Wellesley D., Kurinczuk J.J., Rankin J., de Hoogh K., Fuller G.W, Elliot P., and Toledano M.B. “Risk of congenital anomalies near municipal waste incinerators in England and Scotland: Retrospective population-based cohort study”. Environment International (Parkes et al).

found no association with congenital abnormalities. Under objective (ii), there was a small excess risk, but the paper's authors note that this may be due to residual confounding.

The Imperial College website includes Frequently Asked Questions on this study. One of these is "Does the study show that MWIs are causing increased congenital anomalies in populations living nearby?" The answer is as follows:

"No. The study does not say that the small excess risks associated with congenital heart disease and genital anomalies in proximity to MWIs are caused by those MWIs, as these results may be explained by residual confounding factors i.e. other influences which it was not possible to take into account in the study. This possible explanation is supported further by the fact that the study found no increased risk in congenital anomalies due to exposure to emissions from incinerators."

These three recent papers consider facilities in the UK, operating under the same regulatory regime which would apply to the Facility and operating to the current standards of the IED. The papers found no conclusive evidence of an association of waste incineration facilities with the health outcomes considered. Given that the gasification plant would actually operate to tighter standards, as it would be subject to the reduced emissions limits from the Waste Incineration BREF, the conclusions are directly relevant and support PHE's position statement that *"any potential damage to the health of those living close-by is likely to be very small, if detectable"*.

Therefore, it can be concluded that the effect of emissions from the Facility of pollutants that accumulate in the environment would not be significant. Nonetheless, a quantitative assessment of the effect of emissions has been undertaken and is presented in this assessment.

3 Issue Identification

3.1 Issue

The key issue for consideration is the release of substances to atmosphere from the gasification plant which have the potential to harm human health. No other local sources or sources within the Facility will include emissions of dioxins or dioxin-like PCBs.

The Facility is located off Charlton Lane, Shepperton, immediately to the east of the M3 with a mixture of fields and residential areas surrounding.

Emission limit values (ELVs) are prescribed in the Industrial Emissions Directive (IED) for pollutants known to be produced during the combustion of waste which have the potential to impact upon the local environment either on human health or ecological receptors. A Dispersion Modelling Assessment has been undertaken to support the EP application, which assesses the impact of inhalation of these pollutants on human health. However, dioxins and dioxin-like PCBs can accumulate in the environment, which means that inhalation is only one of the potential exposure routes. Therefore, impacts cannot be evaluated in terms of their effects on human health by simply reference to ambient air quality standards. An assessment needs to be made of the overall human exposure to the substances by the local population and the risk that this exposure causes. The gasification plant has been designed to meet the ELVs outlined in the existing EP but will be required to achieve the BAT-Associated Emission Levels (BAT-AELs) by the end of 2023. As this assessment is associated with the long-term impact of dioxins it has been assumed that the gasification plant operates at the BAT-AEL rather than the ELV in the existing EP.

3.2 Chemicals of Potential Concern (COPC)

The following substances have been considered COPCs for the purpose of this assessment:

- PCDD/Fs (individual congeners); and
- Dioxin-like PCBs;

This assessment investigates the potential for long term health effect of these COPCs through other routes than just inhalation.

4 Assessment Criteria

IRAP calculates the total exposure through each of the different pathways so that a dose from inhalation and ingestion can be calculated for each receptor. By default, these doses are then used to calculate a cancer risk, using the USEPA's approach. However, the Environment Agency recommends that the results be assessed using the UK's approach, which is explained in the Environment Agency's document "Human Health Toxicological Assessment of Contaminants in Soil", ref SC050021.

For the COPCs considered, which have a threshold level for toxicity, a Tolerable Daily Intake (TDI) is defined. This is "an estimate of the amount of a contaminant, expressed on a bodyweight basis, which can be ingested daily over a lifetime without appreciable health risk." A Mean Daily Intake (MDI) is also defined, which is the typical intake from background sources (including dietary intake) across the UK. In order to assess the impact of the gasification plant, the predicted intake of a substance due to emissions from the gasification plant is added to the MDI and compared with the TDI.

Table 1 outlines the MDIs (the typical intake from existing background sources) and TDIs for dioxins and dioxin-like PCBs. These figures are defined in the "Contaminants in soil: updated collation of toxicology data and intake values for humans" series of toxicological reports, available from the UK Government website.

Table 1: Intake of Dioxins and Dioxin-Like PCBs

Item	Units	Intake	
		70 kg adult	20 kg child
Tolerable Daily Intake (TDI)	pg WHO-TEQ/kg bw/day	2.0	
Mean Daily Intake (MDI)	pg WHO-TEQ/kg bw/day	0.7	1.8
	% of TDI	35.00%	90.65%

To allow comparison with the TDI for dioxins, intake values for each dioxin are multiplied by a factor known as the WHO-TEF. A full list of the WHO-TEF values for each dioxin is provided in Table 7.

The TDI has been set at a level which can be ingested daily over a lifetime without appreciable health risk. Therefore, if the total exposure is less than the TDI, it can be concluded that the impact of the Facility is not significant.

5 Conceptual Site Model

5.1 Conceptual site model

A detailed Human Health Risk Assessment has been carried out using the Industrial Risk Assessment Program-Human Health (IRAP-h View – Version 5.0). The programme, created by Lakes Environmental, is based on the United States Environment Protection Agency (USEPA) Human Health Risk Assessment Protocol for Hazardous Waste Combustion Facilities⁵. This Protocol is a development of the approach defined by Her Majesties Inspectorate on Pollution (HMIP) in the UK in 1996⁶, taking account of further research since that date. The exposure pathways included in the IRAP model are shown in Table 2.

Exposure to gaseous contaminants has the potential to occur by direct inhalation or vapour phase transfer to plants. In addition, exposure to particulate phase contaminants may occur via indirect pathways following the deposition of particles to soil. These pathways include:

- ingestion of soil and dust;
- uptake of contaminants from soil into the food-chain (through home-grown produce and crops); and
- direct deposition of particles onto above ground crops.

The pathways through which inhalation and ingestion occur and the receptors that have been considered to be impacted via each pathway are shown in Table 2.

Table 2: Pathways Considered

Pathway	Residential	Allotment	Agricultural
Direct inhalation	Yes	Yes	Yes
Ingestion of soil	Yes	Yes	Yes
Ingestion of home-grown produce	Yes	Yes	Yes
Ingestion of drinking water	Yes	Yes	Yes
Ingestion of eggs from home-grown chickens	-	Yes	Yes
Ingestion of home-grown poultry	-	Yes	Yes
Ingestion of home-grown beef	-	-	Yes
Ingestion of home-grown pork	-	-	Yes
Ingestion of home-grown milk	-	-	Yes
Ingestion of breast milk (infants only)	Infants only		

An agricultural receptor is one with full agricultural interest, typically a commercial farm. A residential receptor is a standard dwelling or school. The allotment type receptor is representative of ingestion pathways for individuals who grow produce, keep chickens and consume eggs and potentially the birds.

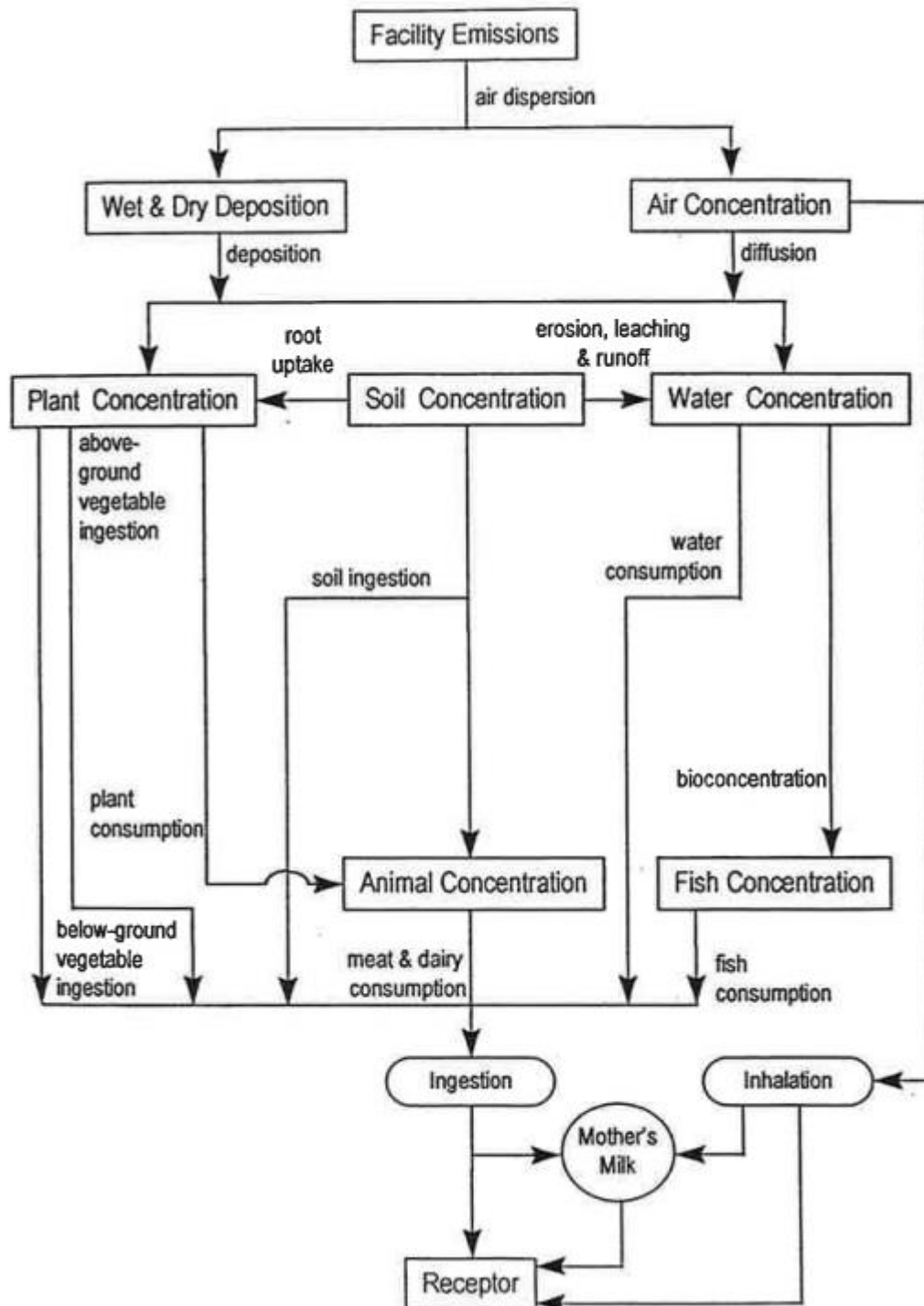
As shown in Figure 1, the pathway from the ingestion of mother's milk in infants is considered within the assessment. The IRAP model calculates the amount of dioxins entering the mother's milk

⁵ USEPA (2005) Human Health Risk Assessment Protocol for Hazardous Waste Combustion Facilities.

⁶ HMIP (1996) Risk Assessment of Dioxin Releases from Municipal Waste Incineration Processes.

and being passed on to the infants. IRAP does not include data on individual PCBs, but it does include data for take-up and accumulation rates within the food chain for two groups of PCBs, known as Aroclor 1254 and Aroclor 1016. IRAP does not include these when determining the intake via mother's milk. Therefore, a safety factor of 1.5 has been applied to the dioxin and dioxin-like PCBs emission rate when considering the impact of the intake via mother's milk. The impacts are then compared against the TDI.

Figure 1: Conceptual Site Model – Exposure Pathways



5.2 Pathways excluded from assessment

The intake of dioxins via dermal absorption, groundwater and surface water exposure and fish consumption pathways is very limited and as such these pathways are excluded from this assessment. The justification for excluding these pathways is highlighted in sections 5.2.1 to 5.2.4.

5.2.1 Dermal absorption

Both the HMIP and the USEPA note that the contribution from dermal exposure to soils impacted from thermal treatment facilities is typically a very minor pathway and is typically very small relative to contributions resulting from exposures via the food chain.

The USEPA⁷ provide an example from the risk assessment conducted for the Waste Technologies, Inc. hazardous thermal treatment in East Liverpool, Ohio. This indicated that for an adult subsistence farmer in a subarea with high exposures, the risk resulting from soil ingestion and dermal contact was 50-fold less than the risk from any other pathway and 300-fold less than the total estimated risk.

The HMIP document⁸ provides a screening calculation using conservative assumptions, which states that the intake via dermal absorption is 30 times lower than the intake via inhalation, which is itself a minor contributor to the total risk.

As such the pathway from dermal absorption is deemed to be an insignificant risk and has been excluded from this assessment.

5.2.2 Groundwater

Exposure via groundwater can only occur if the groundwater is contaminated and consumed untreated by an individual.

The USEPA⁹ have concluded that the build-up of dioxins in the aquifer over realistic travel times relevant to human exposure was predicted to be so small as to be essentially zero.

As such the pathway from groundwater is deemed to be an insignificant risk and has been excluded from this assessment.

5.2.3 Surface water

A possible pathway is via deposition of emissions directly onto surface water – i.e. local drinking water supplies or rainwater storage tanks.

Surface water generally goes through several treatment steps and as such any contaminants would be removed from the water before consumption. Run off to rainwater tanks may not go through the same treatment. However, rainwater tanks have a very small surface area and as such the potential for deposition and build-up of COPCs is limited. As such, the pathway from contaminated surface water is deemed to be an insignificant risk and has been excluded from this assessment.

⁷ USEPA (2005) Human Health Risk Assessment Protocol for Hazardous Waste Combustion Facilities.

⁸ HMIP (1996) Risk Assessment of Dioxin Releases from Municipal Waste Incineration Processes.

⁹ USEPA (2005) Human Health Risk Assessment Protocol for Hazardous Waste Combustion Facilities.

5.2.4 Fish consumption

The consumption of locally caught fish has been excluded from the assessment. Whilst fish makes up a proportion of the UK diet, it is not likely that this would be sourced wide-scale from close proximity to the Facility.

A review of the local waterbodies has been undertaken to see if there are any game fishing lakes in the local area¹⁰. Two have been identified within 3 km of the Facility:

- Ashmere Trout Fishery approximately 1.6 km to the south-east of the Facility; and
- Halliford Mere Trout Fishery approximately 2.2 km to the south of the Facility.

Whilst trout from these fisheries could potentially be consumed by humans these are primarily catch and release fisheries which can only be fished a proportion of the year and there are limits on the number of fish which can be taken per person. As such the potential for fish from either of these fisheries making a significant contribution to a persons diet is minimal. Therefore, there is little risk of dioxins entering the food chain via locally sourced fish and this pathway has been excluded from this assessment.

¹⁰Locations Map, <http://www.fisharound.net/where-to-fish/locations-map>

6 Sensitive Receptors

This assessment considers the possible effects on human health at key receptors, where humans are likely to be exposed to the greatest impact from the Facility, and at the point of maximum impact of annual mean emissions.

For the purposes of this assessment, receptor locations have been categorised as ‘residential’, ‘allotment’ or ‘agricultural’. Residential receptors represent a known place of residence that is occupied within the study area. Agricultural receptors represent a farm holding or area of arable or pastoral farming interest.

The specific receptors identified in the Dispersion Modelling Assessment have been considered in this assessment. In addition, a receptor has been assessed at the point of maximum impact, although this point is actually not in a habited or agricultural area. An agricultural receptor at the point of maximum impact represents the theoretical worst-case impact of the Facility.

These sensitive receptors are listed in Table 3. Reference should be made to Appendix B which shows the location of these receptors with respect to the Facility.

Table 3: Sensitive Receptors

ID	Receptor Name	Location		Type of Receptor
		X	Y	
MAX	Point of maximum impact	508734	169100	Agricultural / Allotment/ Residential
R1	Charlton Road South	508091	168086	Resident
R2	Nutty Lane	508155	168575	Resident
R3	Charlton Road North	508264	168770	Resident
R4	Hetherington Road	508483	169179	Resident
R5	Hawthorn Way North	508807	168458	Resident
R6	Hawthorn Way South	508749	168289	Resident
R7	Watersplash Road	507453	168557	Resident
R8	Birch Grove	509019	169047	Resident

7 IRAP Model Assumptions and Inputs

The following section details the user defined assumptions used within the IRAP model and provides justifications where appropriate.

7.1 Concentrations in soil

The concentration of each chemical in the soil is calculated from the deposition results of the air quality modelling for vapour phase and particle phase deposition. The critical variables in calculating the accumulation of pollutants in the soil are as follows:

- the lifetime of the Facility is taken as 30 years; and
- the soil mixing depth is taken as 2 cm in general and 30 cm for produce.

The split between the solid and vapour phase for the substance considered depends on the specific physical properties of each chemical.

In order to assess the amount of substance which is lost from the soil each year through volatilisation, leaching and surface run-off, a soil loss constant is calculated. The rates for leaching and surface runoff are taken as constant, while the rate for volatilisation is calculated from the physical properties of each substance.

7.2 Concentrations in plants

The concentrations in plants are determined by considering direct deposition and air-to-plant transfer for above ground produce, and root uptake for above ground and below ground produce.

The calculation takes account of the different types of plant. For example, uptake of substances through the roots will differ for below ground and above ground vegetables, and deposition onto plants will be more significant for above ground vegetables.

7.3 Concentrations in animals

The concentrations in animals are calculated from the concentrations in plants, assumed consumption rates and bio-concentration factors. These vary for different animals and different substances, since the transfer of chemicals between the plants consumed and animal tissue varies.

It is also assumed that 100% of the plant materials eaten by animals is grown on soil contaminated by emission sources. This is likely to be a highly pessimistic assumption for UK farming practice.

7.4 Concentrations in humans

7.4.1 Intake via inhalation

This is calculated from inhalation rates of typical adults and children and atmospheric concentrations. The inhalation rates used for adults and children are:

- adults – 20 m³/day; and
- children – 7.2 m³/day.

These are as specified within the Environment Agency's series of reports: "Contaminants in soil: updated collation of toxicology data and intake values for humans". The calculation also takes account of time spent outside, since most people spend most of their time indoors.

7.4.2 Intake via soil ingestion

This calculation allows for the ingestion of soil and takes account of different exposure frequencies. It allows for ingestion of soil attached to unwashed vegetables, unintended ingestion when farming or gardening and, for children, ingestion of soil when playing.

7.4.3 Ingestion of food

The calculation of exposure due to ingestion of food draws on the calculations of concentrations in animals and plants and takes account of different ingestion rates for the various food groups by different age groups.

For most people, locally-produced food is only a fraction of their diet and so exposure factors are applied to allow for this.

7.4.4 Breast milk ingestion

For infants, the primary route of exposure is through breast milk. The calculation draws on the exposure calculation for adults and then allows for the transfer of chemicals in breast milk to an infant who is exclusively breastfed.

The only pathway considered for dioxins for a breastfeeding infant is through breast milk. The modelled scenario consists of the accumulation of pollutants in the food chain up to an adult receptor, the accumulation of pollutants in breast milk and finally the consumption of breast milk by an infant.

The assumptions used were:

- | | |
|--|--------------|
| • Exposure duration of infant to breast milk | 1 year |
| • Proportion of ingested dioxin that is stored in fat | 0.9 |
| • Proportion of mother's weight that is stored in fat | 0.3 |
| • Fraction of fat in breast milk | 0.04 |
| • Fraction of ingested contaminant that is absorbed | 0.9 |
| • Half-life of dioxins in adults | 2,555 days |
| • Ingestion rate of breast milk | 0.688 kg/day |
| • Safety factor on total breast milk dioxin intake to account for PCBs | 1.5 |

7.5 Estimation of COPC concentration in media

The IRAP-h model uses a database of physical and chemical parameters to calculate the COPC concentrations through each of the different pathways identified. The base physical and chemical parameters have been used in this assessment.

In order to calculate the COPC concentrations, a number of site-specific pieces of information are required.

Weather data was obtained for the period 2017 to 2021 from the Heathrow Airport weather station, as used within the dispersion modelling. This provides the annual average precipitation which can be used to calculate the general IRAP-h input parameters, as presented in Table 4.

Table 4: *Ground Type Dependent Properties*

Input Variable	Assumption	Value (cm/year)
Annual average evapotranspiration	70% of annual average precipitation	43.08
Annual average irrigation	0% of annual average precipitation	0.00
Annual average precipitation	100% of annual average precipitation	61.54
Annual average runoff	10% of annual average precipitation	6.15

The average wind speed was taken as 3.94 m/s, calculated from the average of the five years of weather data from Heathrow Airport.

A number of assumptions have been made with regard to the deposition of the different phases. These are summarised in Table 5.

Table 5: *Deposition Assumptions*

Deposition Phase	Dry Deposition Velocities (m/s)	Ratio Dry deposition to Wet deposition	
		Dry Deposition	Wet Deposition
Vapour	0.005	1.0	2.0
Particle	0.010	1.0	2.0
Bound particle	0.010	1.0	2.0
Mercury vapour	0.029	1.0	0.0

Note: the above deposition velocities have been agreed with the UK Environment Agency for all IRAP based assessments where modelling of specific deposition of pollutants is not undertaken. These are considered to be conservative.

These deposition assumptions have been applied to the annual mean concentrations predicted using the dispersion modelling, to generate the inputs needed for the IRAP modelling. For details of the dispersion modelling methodology please refer to the Dispersion Modelling Assessment submitted with the EP application.

7.6 Modelled emissions

For the purpose of this assessment, it is assumed that the gasification operates at the BAT-AEL for its entire operational life. In reality, the gasification plant will be shut down for periods of planned maintenance every year, and will typically operate below the BAT-AEL to demonstrate compliance.

The following tables present the emissions rates of each COPC modelled and the BAT-AELs which have been used to derive the emission rate.

Table 6: COPC Emissions Modelled

COPC	Split of congeners for a release of 1 ng I-TEQ/Nm ³ (1)	Emission conc. (ng/Nm ³) ⁽²⁾	Emission rate (ng/s) ⁽³⁾
Sum I-TEQ dioxins ⁽⁴⁾	-	0.06 ng I-TEQ/Nm ³	-
2,3,7,8-TCDD	0.031	0.002	0.021
1,2,3,7,8-PeCDD	0.245	0.015	0.164
1,2,3,4,7,8-HxCDD	0.287	0.017	0.192
1,2,3,6,7,8-HxCDD	0.258	0.015	0.173
1,2,3,7,8,9-HxCDD	0.205	0.012	0.137
1,2,3,4,6,7,8-HpCDD	1.704	0.102	1.140
OCDD	4.042	0.242	2.703
2,3,7,8-TCDF	0.277	0.017	0.185
1,2,3,7,8-PCDF	0.277	0.017	0.185
2,3,4,7,8-PCDF	0.535	0.032	0.358
1,2,3,4,7,8-HxCDF	2.179	0.131	1.457
1,2,3,6,7,8-HxCDF	0.807	0.048	0.540
1,2,3,7,8,9-HxCDF	0.042	0.003	0.028
2,3,4,6,7,8-HxCDF	0.871	0.052	0.583
1,2,3,4,6,7,8-HpCDF	4.395	0.264	2.939
1,2,3,4,7,8,9-HpCDF	0.429	0.026	0.287
OCDF	3.566	0.214	2.385
Total	20.150	1.209	13.476
Dioxin-like PCBs	-	0.103	0.103

Notes:

(1) Split of the congeners taken from Table 7.2a from the HMIP document.

(2) All emissions are expressed at reference conditions of dry gas, 11% oxygen, 273.15K.

(3) Emission release rate calculated by multiplying the normalised volumetric flow rate by the emission concentration.

(4) The WI BREF includes an emission limit for dioxins of 0.06 ng I-TEQ/Nm³, or a combined limit of 0.08 ng I-TEQ/Nm for dioxins when dioxin-like PCBs are included for existing plants. As this assessment considers dioxin-like PCBs separately, the lower limit of 0.06 ng I-TEQ/Nm³ for dioxins has been used.

A number of points should be noted for the two groups of COPCs:

1. Dioxins

The split of the different dioxins and furans is based on split of congeners for a release of 1 ng I-TEQ/Nm³ as presented in in Table 6. This data is taken from Table 7.2a from the HMIP document “Risk Assessment of Dioxin Releases from Municipal Waste Incineration Processes”.

To determine the emission rates, this split of the different dioxins has been multiplied by normalised volumetric flow rate to determine the release rate of each congener.

2. Dioxin-like PCBs

There are a total of 209 PCBs, which act in a similar manner to dioxins, are generally found in complex mixtures and also have TEFs.

The UK Environment Agency has advised that 44 measurements of dioxin like PCBs have been taken at 24 MWIs between 2008 and 2010. The following data summarises the measurements, all at 11% reference oxygen content:

- Maximum = 9.2×10^{-3} ng[TEQ]/m³
- Mean = 2.6×10^{-3} ng[TEQ]/m³
- Minimum = 5.6×10^{-5} ng[TEQ]/m³

For the purpose of this assessment, the maximum monitored PCB concentration has been used which has been converted to an emission rate using the volumetric flow.

The IRAP software, and the HHRAP database which underpins it, does not include any data on individual PCBs, but it does include data for take-up and accumulation rates within the food chain for two groups of PCBs, known as Aroclor 1254 and Aroclor 1016. Each Aroclor is based on a fixed composition of PCBs. Since we are not aware of any data on the specification of PCBs within incinerator or co-incinerator emissions, as a worst-case assumption it has been assumed that PCB emissions consist entirely of each of the two Aroclor compositions and the maximum impact of either composition has been presented.

As shown in Table 1, the MDI and TDI for dioxins and dioxin-like PCBs is given in pg WHO-TEQ/kg bw/day. However, the split of congeners shown in Table 6 which are used to calculate the release rate of each dioxin are based on the I-TEFs listed in Annex VI Part II of the IED. To determine the total intake TEQ for comparison with the TDI, the output of the IRAP model has been multiplied by the relevant WHO-TEFs. The I-TEFs and WHO-TEFs are shown in Table 7.

Table 7: Toxic Equivalency Factors for Dioxins and Furans

Congener	IED I-TEQ Multiplier	2005 WHO-TEF Multiplier
2,3,7,8-TCDD	1	1
1,2,3,7,8-PeCDD	0.5	1
1,2,3,4,7,8-HxCDD	0.1	0.1
1,2,3,6,7,8-HxCDD	0.1	0.1
1,2,3,7,8,9-HxCDD	0.1	0.1
1,2,3,4,6,7,8-HpCDD	0.01	0.01
OCDD	0.001	0.0003
2,3,7,8-TCDF	0.1	0.1
1,2,3,7,8-PCDF	0.5	0.03
2,3,4,7,8-PCDF	0.05	0.3
1,2,3,4,7,8-HxCDF	0.1	0.1
1,2,3,6,7,8-HxCDF	0.1	0.1
1,2,3,7,8,9-HxCDF	0.1	0.1
2,3,4,6,7,8-HxCDF	0.1	0.1
1,2,3,4,6,7,8-HpCDF	0.01	0.01

Congener	IED I-TEQ Multiplier	2005 WHO-TEF Multiplier
1,2,3,4,7,8,9-HpCDF	0.01	0.01
OCDF	0.001	0.0003

Source: Contaminants in soil: updated collation of toxicological data and intake values for humans, Dioxins, furans and dioxin-like PCBs (Science report: SC050021/TOX 12), Environment Agency, 2009

8 Results

8.1 Assessment against TDI - point of maximum impact

Table 8 presents the impact of emissions of dioxins and dioxin-like PCBs from the Facility at the point of maximum impact for an 'Agricultural' receptor. As explained in section 4, this receptor type assumes the direct inhalation, and ingestion from soil, drinking water, and home-grown eggs and meat, beef, pork, and milk. This assumes that the person lives at the point of maximum impact and consumes home-grown produce etc. This is a very worst-case scenario, especially given that the point on maximum impact is uninhabited and not located on cultivated or pastoral land. Reference should be made to Appendix B for the location of the point in relation to the Facility.

Table 8: Impact Analysis – Dioxins and Dioxin-Like PCBs – Point of Maximum Impact

Receptor Type	MDI (% of TDI)	Process Contribution (% of TDI)	Overall (% of TDI)
Adult			
Agricultural	35.00%	1.20%	36.20%
Allotment	35.00%	0.07%	35.07%
Residential	35.00%	0.03%	35.03%
Child			
Agricultural	90.65%	1.71%	92.36%
Allotment	90.65%	0.14%	90.79%
Residential	90.65%	0.10%	90.75%

The TDI is an estimate of the amount of a contaminant, expressed on a bodyweight basis, which can be ingested daily over a lifetime without appreciable health risk. As shown in Table 8, for the worst-case receptor the overall impact (including the contribution from existing dietary intakes) is less than the TDI for dioxins and dioxin-like PCBs. Therefore, there would not be an appreciable health risk based on the emission of these pollutants.

8.2 Breast milk exposure

The total accumulation of dioxins in an infant due to emissions from the Facility, considering the breast milk pathway and based on an adult agricultural receptor at the point of maximum impact feeding an infant, is 0.206 pg WHO-TEQ / kg-bw / day which is 10.31% of the TDI. This is only 0.59% and 0.20% of the TDI for an allotment and residential-type receptor respectively. There are no ingestion pathways besides breast milk ingestion for an infant receptor. As the process contribution is less than the TDI, it is considered that the Facility will not increase the health risks from the accumulation of dioxins in infants significantly.

8.3 Maximum impact at a receptor

Table 9 outlines the impact of emissions from the Facility at the most affected receptor (i.e. the receptor with the greatest impact from ingestion and inhalation of emissions) (R8 – Birch Grove). This has been classified as a residential receptor.

Table 9: Dioxins and Dioxin-Like PCBs Impact – Maximum Impacted Receptor (R8)

Receptor Type	MDI (% of TDI)	Process Contribution (% of TDI)	Overall (% of TDI)
Adult			
Residential	35.00%	0.02%	35.02%
Child			
Residential	90.65%	0.07%	90.72%

As shown, for the most impacted receptor the overall impact (including the contribution from existing dietary intakes) is less than the TDI for dioxins and dioxin-like PCBs. Therefore, there would not be an appreciable health risk based on the emission of these pollutants.

In addition, the total accumulation of dioxins in an infant due to emissions from the Facility, considering the breast milk pathway and based on an adult residential receptor at R8 – Birch Grove feeding an infant, is 0.003 pg WHO-TEQ / kg-bw / day which is 0.14% of the TDI. As the process contribution is less than the TDI, it is considered that the Facility will not increase the health risks from the accumulation of dioxins in infants significantly.

8.4 Uncertainty and sensitivity analysis

To account for uncertainty in the modelling, the impact on human health was assessed for a receptor at the point of maximum impact.

To account for uncertainty in the dietary intake of a person, both residential and agricultural receptors have been assessed. The agricultural receptor is assumed to consume a greater proportion of home grown produce, which has the potential to be contaminated by the COPCs released, than for a residential receptor. In addition, the agricultural receptor includes the pathway from consuming animals grazed on land contaminated by the emission source. This assumes that 100% of the plant materials eaten by the animals is grown on soil contaminated by emission sources.

The agricultural receptor at the point of maximum impact is considered the upper maximum of the impact of the Facility, noting that this is an overestimation of impacts as the land is not used for crop or dairy production.

The IRAP software, and the HHRAP database which underpins it, does not include any data on individual PCBs, but it does include data for take-up and accumulation rates within the food chain for two groups of PCBs, known as Aroclor 1254 and Aroclor 1016. Each Aroclor is based on a fixed composition of PCBs. Since we are not aware of any data on the specification of PCBs within incinerator or co-incinerator emissions, it has been assumed that PCB emissions consist entirely of each of the two Aroclor compositions and the maximum impact of either composition has been presented.

IRAP does not include these Aroclors (which are being used as a proxy for dioxin-like PCBs) when determining the intake via mother's milk. Therefore, a safety factor of 1.5 has been applied to dioxins and dioxin-like PCBs on when considering the impact of the intake via mother's milk. This is slightly more conservative than the emission limit of 0.08 ng I-TEQ/Nm³ for dioxins and furans when dioxin-like PCBs are included.

9 Conclusions

This HHRA has been undertaken based on the following conservative assumptions:

- the gasification plant will operate continually at the BAT-AEL, i.e. at the maximum concentrations which it is expected that the gasification plant will be permitted to operate at; and
- the hypothetical maximum impacted receptor (an agricultural receptor at the point of maximum impact) only ingests food and drink sourced from the area with the maximum contribution from the Facility.

The results of the assessment show that, for an agricultural child receptor at the point of maximum impact, the combined intake from the Facility plus the existing MDI intake of dioxins and dioxin-like PCBs via inhalation and ingestion is below the TDI. In addition, the ingestion of dioxins by an infant being breast fed by an agricultural receptor at the point of maximum impact is less than the TDI. The impact at identified receptor locations is even less. Therefore, there would not be an appreciable health risk based on the emission of these pollutants.

In conclusion, the impact of emissions of dioxins and dioxin-like PCBs from the Facility on human health is not predicted to be significant.

Appendices

A Detailed Results Tables

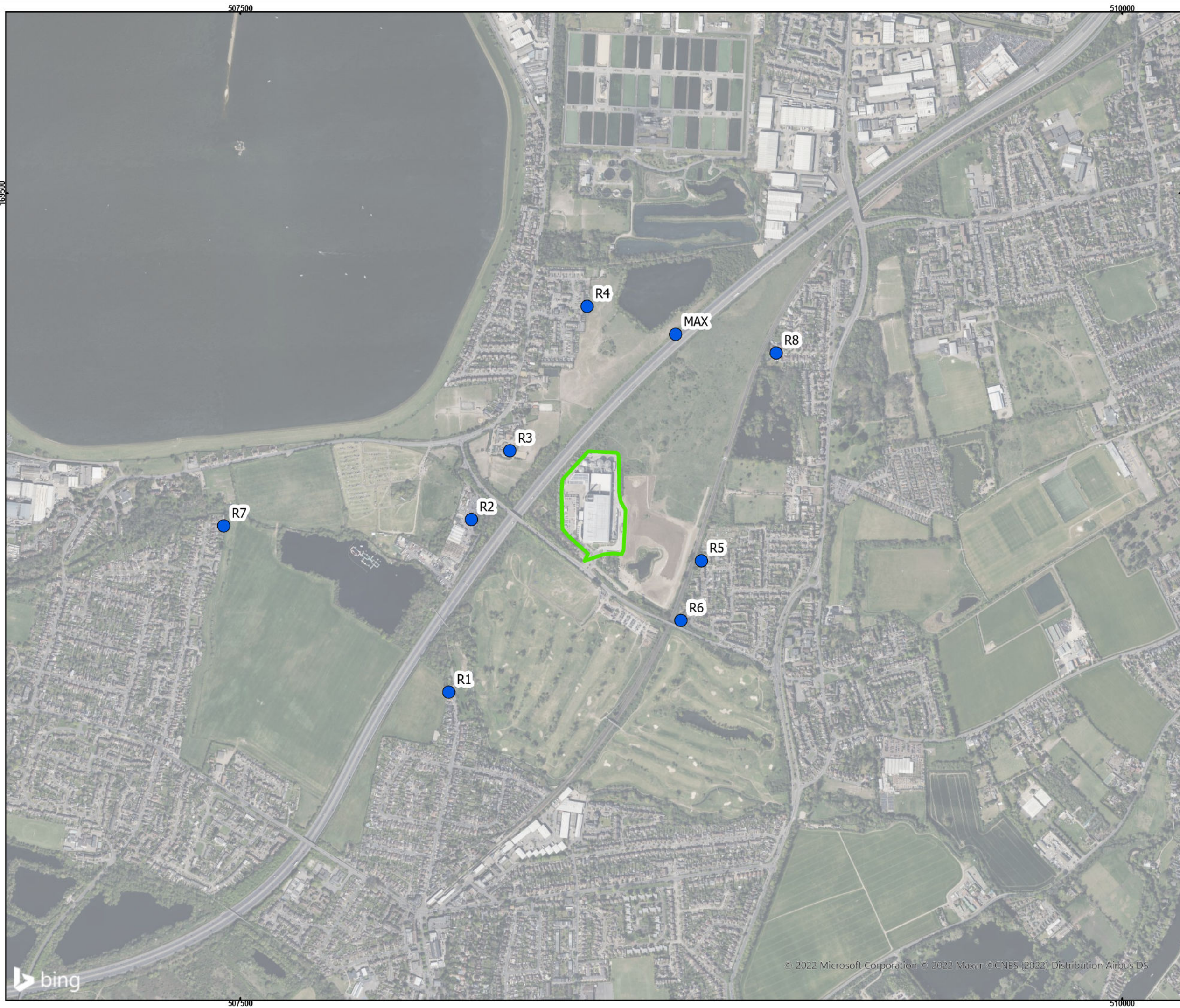
Table 10: Comparison with Total Dioxin and Dioxin-Like PCBs TDI Limits for Adult Receptors

Receptor	Total Inhalation, (pg WHO-TEQ kg ⁻¹ bw day ⁻¹)	Total Ingestion, (pg WHO-TEQ kg ⁻¹ bw day ⁻¹)	Total uptake, (pg WHO-TEQ kg ⁻¹ bw day ⁻¹)	Comparison (% of limit)
MDI (% of TDI)				35.00%
Point of maximum impact - agricultural	6.54E-05	2.39E-02	2.40E-02	36.199%
Point of maximum impact - allotment	6.54E-05	1.30E-03	1.36E-03	35.068%
Point of maximum impact - residential	6.54E-05	4.75E-04	5.41E-04	35.027%
R1 - Charlton Road South	3.35E-05	2.44E-04	2.77E-04	35.014%
R2 - Nutty Lane	2.11E-05	1.53E-04	1.74E-04	35.009%
R3 - Charlton Road North	9.15E-06	6.65E-05	7.56E-05	35.004%
R4 - Hetherington Road	4.08E-05	2.97E-04	3.38E-04	35.017%
R5 - Hawthorn Way North	1.22E-05	8.89E-05	1.01E-04	35.005%
R6 - Hawthorn Way South	1.50E-05	1.09E-04	1.24E-04	35.006%
R7 - Watersplash Road	1.71E-05	1.24E-04	1.41E-04	35.007%
R8 - Birch Grove	4.73E-05	3.43E-04	3.91E-04	35.020%

Table 11: Comparison with Total Dioxin and Dioxin-Like PCBs TDI Limits for Child Receptors

Receptor	Total Inhalation, (pg WHO-TEQ kg ⁻¹ bw day ⁻¹)	Total Ingestion, (pg WHO-TEQ kg ⁻¹ bw day ⁻¹)	Total uptake, (pg WHO-TEQ kg ⁻¹ bw day ⁻¹)	Comparison (% of limit)
MDI (% of TDI)				90.65%
Point of maximum impact - agricultural	8.24E-05	3.41E-02	3.41E-02	92.357%
Point of maximum impact - allotment	8.24E-05	2.68E-03	2.76E-03	90.788%
Point of maximum impact - residential	8.24E-05	1.83E-03	1.92E-03	90.746%
R1 - Charlton Road South	4.22E-05	9.39E-04	9.82E-04	90.699%
R2 - Nutty Lane	2.65E-05	5.91E-04	6.17E-04	90.681%
R3 - Charlton Road North	1.15E-05	2.57E-04	2.68E-04	90.663%
R4 - Hetherington Road	5.15E-05	1.15E-03	1.20E-03	90.710%
R5 - Hawthorn Way North	1.54E-05	3.43E-04	3.58E-04	90.668%
R6 - Hawthorn Way South	1.90E-05	4.22E-04	4.41E-04	90.672%
R7 - Watersplash Road	2.15E-05	4.79E-04	5.01E-04	90.675%
R8 - Birch Grove	5.95E-05	1.32E-03	1.38E-03	90.719%

B Location of Sensitive Receptors



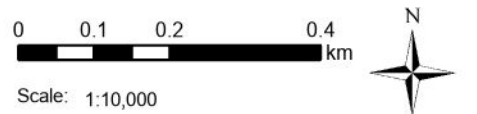
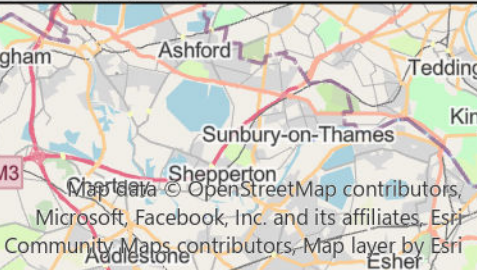
Legend

- Receptor Points
- Installation Boundary

Client:	SUEZ Recycling and Recovery Surrey Ltd
Site:	Charlton Lane Eco Park
Project:	1253
Title:	

DPA - Figure 1. Receptors

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