



Lostock Sustainable Energy Plant Ltd

EIA Report for Proposed Increase to Waste Tonnage Throughput of the Lostock Sustainable Energy Plant

Appendix 5.3 – Vehicle Emissions Modelling



Document approval

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1 Introduction

This appendix contains the detailed methodology and results of the road vehicle modelling carried out to support Chapter 5 - Air Quality and Human Health for the Environmental Impact Assessment (EIA) Report (Main Report), which supports the Proposal to increase the total waste tonnage of the LSEP.

The IAQM document "Land-Use Planning & Development Control: Planning for Air Quality V1.2" (2017), referred to as the IAQM 2017 Planning Guidance, states that an air quality assessment is required where a development would cause a "significant change" in light duty vehicles (LDVs) or heavy duty vehicles (HDV). The indicative criteria to process to an assessment are:

- A change in LDV flows of:
 - o more than 100 Annual Average Daily Traffic (AADT) within or adjacent to an AQMA; or
 - o more than 500 AADT elsewhere.
- A change in HDV flows of:
 - o more than 25 AADT within or adjacent to an AQMA; or
 - o more than 100 AADT elsewhere.

The IAQM guidance does not clearly state the level of assessment which is required. However, if the change in LDV and HDV flows does not exceed the above criteria and there is negligible risk of overlap of emissions with process emissions, the development is not expected to cause a significant change and the significance of effect is deemed to be 'negligible' and further detailed analysis of the impact is not deemed necessary.

As stated in Chapter 4 of the EIA Report (Main Report), the vehicles associated with the operation of the LSEP are expected to result in 514 one-way movements (257 inward journeys and 257 outward journey) on an AADT basis. 434 of these (217 inward journeys and 217 outward journeys) would be HGVs. This is an increase from the number of vehicles within the currents36 consent development of 170 HGV journeys per day and exceeds the criteria above, although there are no AQMAs in the area.

The following table sets out the anticipated vehicle trip generation rate for the operation of the LSEP as set out in Chapter 4 of the EIA Report (Main Report).

Table 1: Traffic Data 24-hour AADT

| Scenario | All vehicles | LDVs | HGVs |
|---------------------------------------|--------------|------|------|
| S36 consent | 324 | 80 | 264 |
| the Proposal (i.e. the increase) | 170 | 0 | 170 |
| Total LSEP vehicles with the Proposal | 514 | 80 | 434 |

In line with the IAQM 2017 Planning Guidance, it is appropriate to assess the impact of the LSEP with the Proposal against the current baseline for the site, disregarding the extant permissions to reflect the real world increase experienced by receptors.

In addition, the routing of vehicles north east along the A556 has the potential for vehicle emissions to combine with process emissions. Therefore, it has been considered appropriate to undertake a detailed assessment of the transport emissions both as they exceed the IAQM screening threshold and in order to calculate the in combination impact with the process emissions.



2 Methodology

2.1 Model used

All traffic modelling was undertaken using the ADMS-Roads (version 5.1) dispersion modelling package. The ADMS-Roads model is a version of ADMS, which was developed by CERC and is commonly used throughout the UK for environmental assessment purposes. ADMS-Roads is routinely used for modelling of emissions for planning purposes to the satisfaction of local authorities.

2.2 Input data

The model requires input data that details the following parameters:

- Traffic flow data;
- Vehicle emission factors;
- Spatial co-ordinates of emissions;
- Discrete receptor points;
- Meteorological data;
- Roughness length; and,
- Monin-Obukhov length.

2.3 Traffic flow data

24-hour AADT flows and HDV numbers have been provided by Axis, the transport consultant for the project, for the following scenarios:

- Scenario 1: 2016 Baseline,
- Scenario 2: 2023 do-minimum: including Northwich Traffic Model growth to represent general traffic growth and committed developments in the area; and
- Scenario 3: 2023 do-something: as scenario 2, plus LSEP traffic flows.

The impact is defined as the difference between the 'do-something' and 'do-minimum' scenarios, i.e. scenario 3 minus scenario 2. The do-minimum excludes the vehicles already approved under the s36 consent as such the impact is the total impact of the vehicles associated with the operation of the LSEP.

The 2016 baseline data provided by Axis has been factored to generate a 2018 baseline data using the Department for Transport (DfT) road traffic statistics from count points along the traffic routeways. This is to allow for the use of the most recent Emissions Factors Toolkit (EFT), version 10.1, which does not support years prior to Version 10.1 of the EFT is more accurate than earlier versions and allows the inclusion of gradient.

The Northwich transport model which has been used to calculate the traffic flows, has been developed by Mott Macdonald on behalf of Cheshire West and Chester County Council (CWaCC). It covers a very wide network, which includes over 80 allocated and committed development sites, the traffic from which was distributed across the whole modelled network. This means that it is not possible to isolate the volume of traffic associated with specific committed developments close to the LSEP, but they have been considered within the model.



The roads included in the model are shown in Figure 1 of Appendix C and the traffic data used in the assessment is presented in Appendix A.

2.3.1 Vehicle speeds

Vehicles have been modelled at speeds in accordance with national speed limits and professional judgments of the roads and vehicles. Speeds have been modelled at 32 kph (20 mph) for LDVs and 27 kph (~17 mph) for HDVs along the site access road; 64 kph (40 mph) for LDVs and 59 kph (~37 mph) for HDVs along Griffiths Road; 113 kph (70 mph) for LDVs and 96 kph (~60 mph) for HDVs along the A556 and A350 to the south of the roundabout; and 48 kph (30 mph) for LDVs and 43 kph (~27 mph) for HDVs along the remainder of the roads.

Slower speeds have been used where appropriate; at all junctions and roundabouts. For smaller roads, a slow down section preceding and after a junction of 25 m has been used, as recommended in Local Air Quality Management Technical Guidance (TG16), DEFRA, 2021). At the larger junctions and roundabouts, a longer slow down distance of 50 m has been used. At the junction of the A556 with Gadbrook Road, these distances have been further increased to ~300 m, to reflect the typical speed of traffic here due to traffic lights. Slow down sections use speeds of 20 kph (~12 mph) have been used for LDVs and HDVs. These sections are shown on Figure 2.

2.3.1.1 Queue zones

A review of typical traffic conditions has been undertaken using Google Maps. This has indicated that queuing occurs and along certain stretches during peak periods, and more generally close to major junctions. Representative queue zones have been modelled. Guidance has been taken from CERC guidance note 60 – Modelling queuing traffic¹. This note recommends the following approach:

- 1. Assume a representative average vehicle length 5.75 m which is the highways industry standard.
- 2. Assume that the vehicles are travelling at the slowest speed it is possible to model (5 kph).
- 3. Calculate a representative AADT for the queue zones. The AADT can be calculated as: $AADT = [speed(m/hour)/vehicle \ length(m] \ x \ 24$
- 4. Using the assumed values from (1) and (2), this gives a representative AADT of 20,870 vehicles.

Emissions from the queue zones have been applied to the hours when queuing is most common based on the review of traffic using Google Maps. The severity of the queuing has also been identified and factored into the model as 25% or 50% as in accordance with the Google Map traffic review. Separate queue profiles have been generated for difference stretches of each of the modelled road links towards and away from the LSEP site.

The emissions from the roads that overlap the queue zone are always on. Whilst this has the potential to over-predict emissions during hours of queuing, it is important to retain these emissions as they will capture the increase in emissions due to development-generated traffic. Queue zones always have the same speed (5 kph) and AADT (20,870 per queue lane), therefore there would be no difference in emissions between scenarios on these road sections for the hours with queuing traffic, unless the emissions from baseline and development traffic were also included. The sections identified as queue zones are highlighted on Figure 2.

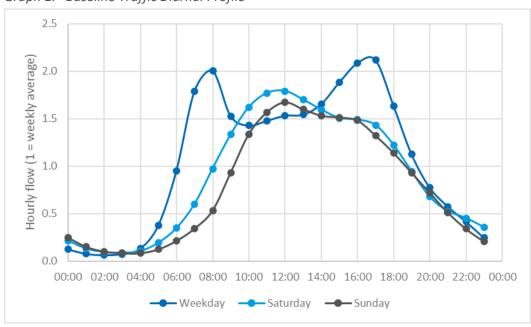
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¹Cambridge Environmental Research Consultants – CERC note 60, Modelling queuing traffic, August 2004



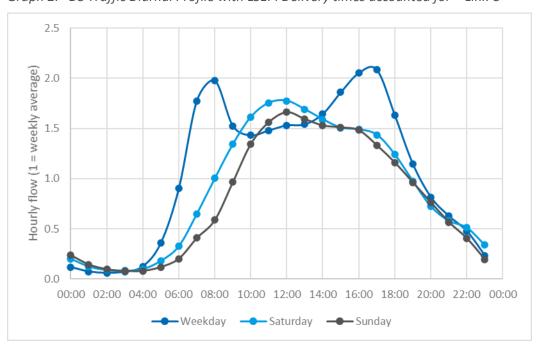
2.4 Daily profile of traffic volume

It is important that the model reflects the variability of the traffic flow during the day. To account for this, a time varying emission profile has been included in the model. This has used the annual daily traffic flow and distribution for 2018 from the road traffic statistics (TRA), provided by the DfT. The time varying profile for the Do Something scenario has also considered additional traffic associated with the operation of the LSEP, and the timings of deliveries between 7 am and 11 pm.



Graph 1: Baseline Traffic Diurnal Profile







2.5 Vehicle emissions factor

Emission factors for NOx, PM_{10} and $PM_{2.5}$ have been determined for each scenario using the traffic data and the EFT v 10.1 (2VC) database of road traffic emission factors within ADMS Roads. All roads were classified as "England (urban)". Emissions for each link have been calculated using the EFT within the ADMS model.

It is possible to account for the effect of road gradients on vehicle emissions using the EFT, although this only affects emissions from HDVs, and the effects are capped at a gradient of 6%. This option has been used where the gradient is expected to significantly influence emissions from HDVs. There are three sections along the A556 (Link Ia and Ib) which have included gradients of 1.7% and 2%.

The EFT predicts that emissions from road vehicles will reduce in future years as newer cleaner vehicles enter the fleet². However, evidence has shown that the rate of this reduction may not be occurring in the real world as the vehicle fleet turnover rate has reduced. Within this assessment, we have conservatively adopted a worse-case scenario, which assumes no change to the fleet composition on local networks between 2018 and the opening year of 2023. The assessment also conservatively applies 2018 background concentrations to the future scenarios – i.e. assumes no reduction in background concentration.

The EFT does not include emissions of ammonia from vehicles. However, petrol vehicles emit ammonia due to the degradation of catalytic converters, and diesel vehicles emit ammonia due to measures to reduce NOx emissions. This has been shown to be a significant source of nitrogen deposition at roadside locations³. Air Quality Consultants (AQC) has published the Calculator for Realistic Emissions of Ammonia (CREAM V1A⁴) for the calculation of emissions of ammonia from vehicles, which has been used to calculate ammonia emissions for each road link and scenario for the assessment of the effect of ammonia on ecological receptors. Ammonia emissions from vehicles are not a concern with regard to human health due to the very high AQAL for the protection of human health and relatively low contributions from vehicles.

2.6 Spatial co-ordinates of vehicle emissions

Street locations and widths were estimated from a desk-top mapping study and referenced to UK National Grid Reference (NGR) co-ordinates.

2.7 Discrete receptor points

2.7.1 Human receptors

The Design Manual for Roads and Bridges (DMRB) considers any receptor within 200 m of a road source to be potentially affected by that operation. The AQALs only apply at locations where the public may be exposed to pollution for a sufficient period for there to be any measurable health effect. Representative receptors have been chosen at the façade of residential properties along the

² In this context 'fleet' refers to the fleet mix, which describes all vehicles on the road, and accounts for a range of vehicle emission standards of LDVs and HGVs

³ Air Quality Consultants, Ammonia Emissions from Roads for Assessing Impacts on Nitrogen-sensitive Habitats, February 2020

⁴ Available from https://www.aqconsultants.co.uk/resources



modelled roads and are shown in Table 2. Reference should be made to Figure 1 which shows the location of each of the discrete receptor locations assessed.

Table 2: Human Sensitive Receptors

| RR1 A556 – Lostock Gralam 369835 375138 1.5 RR2 A556 – Lostock Gralam 369749 374973 1.5 RR3 A556 – Lostock Gralam 369721 374900 1.5 RR4 A556 – Lostock Gralam 369701 374848 1.5 RR5 A556 – Lostock Gralam 369784 374318 1.5 RR6 Fieldhouse Farm 369375 373847 1.5 RR7 A556 – Lostock Green 369375 373847 1.5 RR8 A556 – Lostock Green 369392 373750 1.5 RR10 A556 – Lostock Green 369338 373630 1.5 RR11 A556 – Lostock Green 369338 373500 1.5 RR11 A556 – Lostock Green 369319 373590 1.5 RR11 A556 – Lostock Green 369338 373630 1.5 RR11 A556 – Lostock Green 369339 373286 1.5 RR11 A556 – Lostock Green 369339 373280 <td< th=""><th>ID</th><th>Description</th><th>X (m)</th><th>Y (m)</th><th>Height (m)</th></td<> | ID | Description | X (m) | Y (m) | Height (m) |
|---|------|--------------------------|--------|--------|------------|
| RR3 A556 – Lostock Gralam 369721 374900 1.5 RR4 A556 – Lostock Gralam 369701 374848 1.5 RR5 A556 – Lostock Gralam 369672 374703 1.5 RR6 Fieldhouse Farm 369784 374318 1.5 RR7 A556 – Lostock Green 369375 373847 1.5 RR8 A556 – Lostock Green 369311 373810 1.5 RR10 A556 – Lostock Green 369338 373630 1.5 RR11 A556 – Lostock Green 369319 373590 1.5 RR11 A556 – Lostock Green 369319 373590 1.5 RR12 Cookes Road 369044 373286 1.5 RR12 Cookes Road 369033 3732470 1.5 RR13 Cookes Road 369754 372470 1.5 RR14 High House 368754 372470 1.5 RR14 High House 366754 372479 1.5 RR15< | RR1 | A556 – Lostock Gralam | 369835 | 375138 | 1.5 |
| RR4 A556 – Lostock Gralam 369701 374848 1.5 RR5 A556 – Lostock Gralam 369672 374703 1.5 RR6 Fieldhouse Farm 369784 374318 1.5 RR7 A556 – Lostock Green 369375 373847 1.5 RR8 A556 – Lostock Green 369392 373750 1.5 RR10 A556 – Lostock Green 369338 373630 1.5 RR11 A556 – Lostock Green 369319 373590 1.5 RR12 Cookes Road 369064 373286 1.5 RR13 Cookes Road 369033 373228 1.5 RR14 High House 368736 372470 1.5 RR15 Shurlach Road 367554 372479 1.5 RR16 Shurlach Road 367711 372613 1.5 RR17 Shurlach Road 367789 372659 1.5 RR18 Shurlach Road 367711 372613 1.5 RR19 | RR2 | A556 – Lostock Gralam | 369749 | 374973 | 1.5 |
| RRS A556 – Lostock Gralam 369672 374703 1.5 RR6 Fieldhouse Farm 369784 374318 1.5 RR7 A556 – Lostock Green 369375 373847 1.5 RR8 A556 – Lostock Green 369311 373810 1.5 RR10 A556 – Lostock Green 369328 373750 1.5 RR11 A556 – Lostock Green 369319 373590 1.5 RR12 Cookes Road 369031 373286 1.5 RR13 Cookes Road 369033 373228 1.5 RR14 High House 368736 372470 1.5 RR15 Shurlach Road 367554 372479 1.5 RR16 Shurlach Road 367634 372552 1.5 RR17 Shurlach Road 367711 372613 1.5 RR18 Shurlach Road 367789 372559 1.5 RR19 Elizabethan Way 368351 372712 1.5 RR210 Tudor Close 368415 372712 1.5 RR22 Fo | RR3 | A556 – Lostock Gralam | 369721 | 374900 | 1.5 |
| RR6 Fieldhouse Farm 369784 374318 1.5 RR7 A556 – Lostock Green 369375 373847 1.5 RR8 A556 – Lostock Green 369411 373810 1.5 RR9 A556 – Lostock Green 369392 373750 1.5 RR10 A556 – Lostock Green 369319 373590 1.5 RR11 A556 – Lostock Green 369319 373590 1.5 RR12 Cookes Road 369031 373286 1.5 RR12 Cookes Road 369033 373228 1.5 RR13 Cookes Road 369033 372470 1.5 RR14 High House 368736 372470 1.5 RR14 High House 368736 372470 1.5 RR15 Shurlach Road 36754 372479 1.5 RR16 Shurlach Road 367711 372613 1.5 RR17 Shurlach Road 367711 372613 1.5 RR18 Shurlac | RR4 | A556 – Lostock Gralam | 369701 | 374848 | 1.5 |
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| RR9 A556 – Lostock Green 369392 373750 1.5 RR10 A556 – Lostock Green 369338 373630 1.5 RR11 A556 – Lostock Green 369319 373590 1.5 RR12 Cookes Road 369064 373286 1.5 RR13 Cookes Road 369033 373228 1.5 RR14 High House 368736 372470 1.5 RR15 Shurlach Road 367554 372479 1.5 RR16 Shurlach Road 367634 372552 1.5 RR17 Shurlach Road 367711 372613 1.5 RR18 Shurlach Road 367789 372659 1.5 RR19 Elizabethan Way 368351 372712 1.5 RR19 Elizabethan Way 368351 372712 1.5 RR20 Tudor Close 368415 372719 1.5 RR21 Mulberry Close 368702 372807 1.5 RR22 Foxglove Way 368612 372766 1.5 RR23 Foxglove Way <td>RR7</td> <td>A556 – Lostock Green</td> <td>369375</td> <td>373847</td> <td>1.5</td> | RR7 | A556 – Lostock Green | 369375 | 373847 | 1.5 |
| RR10 A556 – Lostock Green 369338 373630 1.5 RR11 A556 – Lostock Green 369319 373590 1.5 RR12 Cookes Road 369064 373286 1.5 RR13 Cookes Road 369033 373228 1.5 RR14 High House 368736 372470 1.5 RR15 Shurlach Road 367554 372479 1.5 RR16 Shurlach Road 367634 372552 1.5 RR17 Shurlach Road 367711 372613 1.5 RR18 Shurlach Road 367789 372659 1.5 RR19 Elizabethan Way 368351 372712 1.5 RR20 Tudor Close 368415 372719 1.5 RR21 Mulberry Close 368702 372807 1.5 RR22 Foxglove Way 368648 372785 1.5 RR23 Foxglove Way 368648 372713 1.5 RR24 Foxglove Way 368548 372713 1.5 RR25 Tudor Close <t< td=""><td>RR8</td><td>A556 – Lostock Green</td><td>369411</td><td>373810</td><td>1.5</td></t<> | RR8 | A556 – Lostock Green | 369411 | 373810 | 1.5 |
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| RR12 Cookes Road 369064 373286 1.5 RR13 Cookes Road 369033 373228 1.5 RR14 High House 368736 372470 1.5 RR15 Shurlach Road 367554 372479 1.5 RR16 Shurlach Road 367634 372552 1.5 RR17 Shurlach Road 367711 372613 1.5 RR18 Shurlach Road 367789 372659 1.5 RR19 Elizabethan Way 368351 372712 1.5 RR20 Tudor Close 368415 372719 1.5 RR21 Mulberry Close 368702 372807 1.5 RR22 Foxglove Way 368648 372785 1.5 RR23 Foxglove Way 368612 372766 1.5 RR24 Foxglove Way 368548 372751 1.5 RR25 Tudor Close 368468 372713 1.5 RR26 Foxglove Way 368525 372764 1.5 RR27 Tudor Close 368469 | RR10 | A556 – Lostock Green | 369338 | 373630 | 1.5 |
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| RR14 High House 368736 372470 1.5 RR15 Shurlach Road 367554 372479 1.5 RR16 Shurlach Road 367634 372552 1.5 RR17 Shurlach Road 367711 372613 1.5 RR18 Shurlach Road 367789 372659 1.5 RR19 Elizabethan Way 368351 372712 1.5 RR20 Tudor Close 368415 372719 1.5 RR21 Mulberry Close 368702 372807 1.5 RR22 Foxglove Way 368648 372785 1.5 RR23 Foxglove Way 368548 372751 1.5 RR24 Foxglove Way 368548 372751 1.5 RR25 Tudor Close 368468 372713 1.5 RR26 Foxglove Way 368525 372764 1.5 RR27 Tudor Close 368469 372781 1.5 RR28 Broken Cross King Street 368462 372883 1.5 RR30 Broken Cross King Street | RR12 | Cookes Road | 369064 | 373286 | 1.5 |
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| | RR33 | Broken Cross King Street | 368402 | 373034 | 1.5 |
| RR35 Broken Cross King Street 368409 373108 1.5 | RR34 | Broken Cross King Street | 368416 | 373051 | 1.5 |
| | RR35 | Broken Cross King Street | 368409 | 373108 | 1.5 |



| ID | Description | X (m) | Y (m) | Height (m) |
|------|---|--------|--------|------------|
| RR36 | Broken Cross King Street | 368392 | 373142 | 1.5 |
| RR37 | Middlewich Road | 368319 | 373210 | 1.5 |
| RR38 | Middlewich Road | 368334 | 373215 | 1.5 |
| RR39 | Middlewich Road | 368317 | 373256 | 1.5 |
| RR40 | Cottage Close | 368295 | 373444 | 1.5 |
| RR41 | Cottage Close | 368296 | 373463 | 1.5 |
| RR42 | Cottage Close | 368294 | 373487 | 1.5 |
| RR43 | Cottage Close | 368303 | 373515 | 1.5 |
| RR44 | Cottage Close | 368295 | 373541 | 1.5 |
| RR45 | Cottage Close | 368301 | 373557 | 1.5 |
| RR46 | Griffiths Road Junction with Manchester Road | 368620 | 374687 | 1.5 |
| RR47 | Griffiths Road Junction with Manchester Road | 368630 | 374705 | 1.5 |
| RR48 | Griffiths Road Junction with Manchester Road | 368640 | 374725 | 1.5 |

This is not an exhaustive list but is a selection of receptors along each of the road links where vehicles associated with the operation of the LSEP are expected to travel. The human receptors used for the assessment of process emissions have also been assessed within this vehicle emissions assessment – see Table 5 of Appendix 5.2. Only the results of those receptors within 200 m of the roads used by LSEP vehicles (shown in orange stars on Figure 1) have been presented.

2.7.2 Ecological receptors

Four of the ecological receptors identified within the process emissions assessment are located within 200 m of the vehicle routes;

- Wade Brook LWS;
- Long Wood LWS;
- Rudheath Lime Beds pLWS; and
- Winnington Wood LWS and ancient woodland.

To assess the impact at these receptors, a series of transects at right angles to the road, with points every 1 m have been modelled. Due to the extent of Rudheath Lime Beds, three transects have been modelled at this site, and two at Winnington Woods, on either side of the road. Reference should be made to Figure 1 which shows the location of each of the transects assessed.

2.8 Meteorological data

To calculate pollutant concentrations at identified receptor locations, the model uses sequential hourly meteorological data, including wind direction, wind speed, temperature, cloud cover and stability, which exert significant influence over atmospheric dispersion.

Sequential 1-hour meteorological data to be used in this assessment were taken from Manchester Airport meteorological station for 2018. As stated within Appendix 5.2 of the EIA Report, Manchester is considered to be the most representative meteorological station available.

Typically, road assessments use one-year of meteorological data. The traffic baseline data and meteorological data are all for the year 2018. A wind rose of the 2018 meteorological data is provided in Figure 3 of Appendix 5.2 to the EIA Report.

The surface roughness and Monin-Obukov lengths used within the roads modelling for the site location and meteorological site location are the same as those used for the process emissions modelling as justified in Appendix 5.2 of the EIA Report. A constant surface roughness value has been used as it is not possible to use a variable surface roughness file within ADMS Roads.

Table 3: Meteorological Parameters Summary

| Parameter | Dispersion site | Meteorological site |
|------------------------------|-----------------|---------------------|
| Minimum Monin-Obukhov length | 10 m | 10 m |
| Surface roughness | 0.5 m | 0.3 m |

2.9 Background data

For the purpose of this analysis the mapped background concentrations for each receptor point have been extracted from the DEFRA UK AIR background mapping database, for nitrogen dioxide and PM_{10} and $PM_{2.5}$. This data is presented in Table 4. There is uncertainty as to how background pollutant concentrations will change in the future, so as a conservative measure the 2018 background pollutant concentrations have been applied to the future year (2023) scenarios – i.e. assuming no reduction in background pollutant concentrations.

Table 4: Mapped Background Data

| Grid square | Annual mean concentration (μg/m³) | | |
|------------------|-----------------------------------|------------------|-------------------|
| | Nitrogen dioxide | PM ₁₀ | PM _{2.5} |
| Annual mean AQAL | 40 | 18 | 10 |
| 368500, 374500 | 11.45 | 10.64 | 7.04 |
| 368500, 372500 | 11.76 | 11.04 | 7.19 |
| 366500,373500 | 11.22 | 11.17 | 7.53 |
| 367500,372500 | 13.34 | 11.50 | 7.53 |
| 369500,375500 | 10.33 | 11.13 | 7.13 |
| 369500,374500 | 11.42 | 11.38 | 7.29 |
| 369500,373500 | 10.50 | 11.27 | 7.14 |
| 368500,373500 | 10.70 | 10.83 | 7.16 |
| 370500,375500 | 10.90 | 11.36 | 7.16 |

Source: DEFRA 2018 mapped background datasets

As shown the mapped background concentrations are well below the AQAL for nitrogen dioxide, PM_{10} and $PM_{2.5}$.

2.10 Post modelling conversion from NOx to nitrogen dioxide (NO₂)

The modelled road-NOx and the mapped background concentrations have been used as inputs in DEFRA's NOx to NO2 calculator (V8.1) to convert modelled NOx to NO2 in accordance with the methodology outlined in LAQM.TG(16).



When converting from NOx to NO₂ the following inputs have been used:

- The year has been taken as the same as the emissions data, i.e. 2018 for the wort case scenario;
- The local authority has been selected as "Cheshire West and Chester"; and
- The traffic mix has been selected as "All other urban UK traffic".

2.11 Verification

The ADMS Roads model has been validated against real world monitoring. However LAQM.TG(16) recommends that the model output is verified. The verification process should involve the comparison between predicted and measured concentrations at one or more suitable local sites and forms an essential component of a detailed assessment for road traffic models. Part of the verification process involves improvements to the base model to provide a better representation of the monitored data. This includes checks on:

- Traffic data;
- Road widths;
- Distance between sources and monitoring locations;
- Speed estimates;
- Street canyons;
- Background concentrations; and
- Monitoring data.

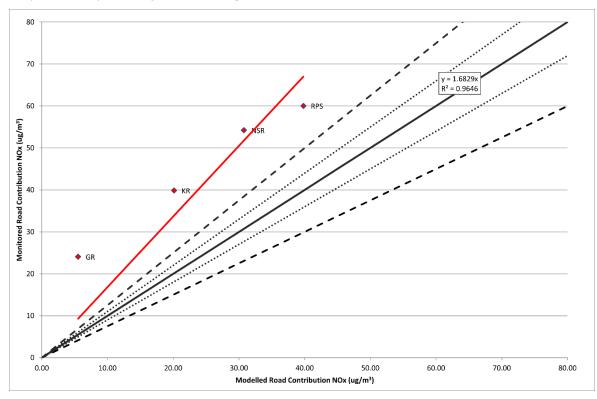
All of these have been reviewed and the model refined to increase the accuracy as much as possible.

LAQM.TG(16) recommends that a number of points are used and the results plotted. The regression factor should then be used as the verification factor. Analysis of a number of data points can be used to see if the model is not performing well in a given area and highlight issues within the modelling such as incorrect traffic data.

There are four monitoring sites with data for 2018 which are situated along the link routes used in the model; Griffiths Road (GR), King St Rudheath (KR), Station Road (NSR) and Rudheath Primary School (RPS). These are all roadside sites.

An initial screening of the sites has identified that both GR and RPS may be difficult to use for verification. This is because both monitoring sites are at junctions with another road which traffic data is not included within the model. GR is at the junction of Griffiths Road and Manchester Road, but only the former of which is included within the model. Therefore, it is likely that the model will underpredict at this point because it does not include the contribution from Manchester Road. This is a similar situation at RPS, for which the model does not include the influence from Gadbrook Road. In addition RPS is situated close to a school carpark which may not be represented in the model.

Graph 3 below shows the unadjusted results at all four of the monitoring sites. It shows that the model is underpredicting at all locations. As expected, the results at GR are very much underpredicting because the model does not include influence from Manchester Road. At RPS, the model is underpredicted but not to the extent expected. This is likely to be because the major junction and queues modelled on the A556 are the greatest contribution to the monitoring site. It is therefore considered appropriate to use RPS within the verification. Graph 4 shows the verification results excluding monitoring point GR. This gives a regression correction factor of 1.6551, confirming that the model is under-predicting road- NOx. The R2 value (a measure of the fit of the data points to the trendline, with a maximum value of 1.00) is 0.9892.



Graph 3: Comparison of Monitored against Modelled Road NOx



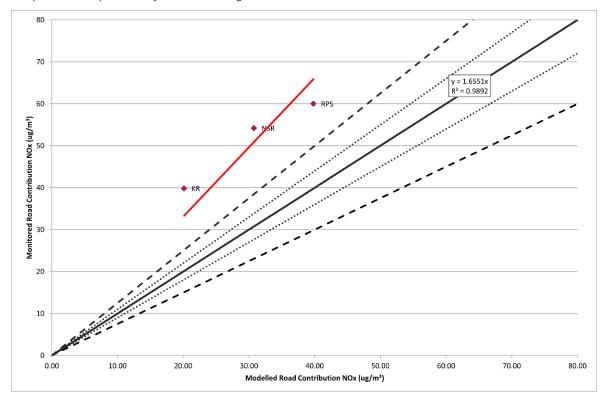


Table 5 details the monitoring locations suitable for verification. In the first instance the monitored road-NOx contribution at each monitoring location has been calculated.

Table 5: Verification Procedure: Monitored Road NOx

| Location | 2018 monitored NO ₂ (μg/m³) | Background NO ₂ (μg/m³) | 2018 calculated road NOx (μg/m³) |
|----------|--|------------------------------------|-------------------------------------|
| KR | 32 | 11.76 | 39.84 |
| NSR | 38 | 11.22 | 54.21 |
| RPS | 42.4 | 13.34 | 60.02 |

Note:

All NOx to NO_2 conversions undertaken using DEFRA's NOx to NO_2 calculator V8.1, for 2018 emissions and using the 'All other urban UK traffic' traffic mix setting.

The modelled road-NOx output has then been compared to the calculated road-NOx concentration, and the modelled total NO_2 compared to the monitored NO_2 concentration.

Table 6: Verification Procedure: Raw Model Results Comparison

| Location | 2018 modelled road NOx (μg/m³) | Ratio of monitored to modelled road NOx | 2018 modelled total NO ₂ (μg/m³) | Ratio of monitored to modelled total NO ₂ |
|----------|--------------------------------------|---|---|--|
| KR | 20.12 | 2.0 | 22.42 | 0.7 |
| NSR | 30.76 | 1.8 | 27.18 | 0.7 |
| RPS | 39.80 | 1.5 | 33.44 | 0.8 |

Note:

All NOx to NO_2 conversions undertaken using DEFRA's NOx to NO_2 calculator V8.1, for 2018 emissions and using the 'All other urban UK traffic' traffic mix setting.

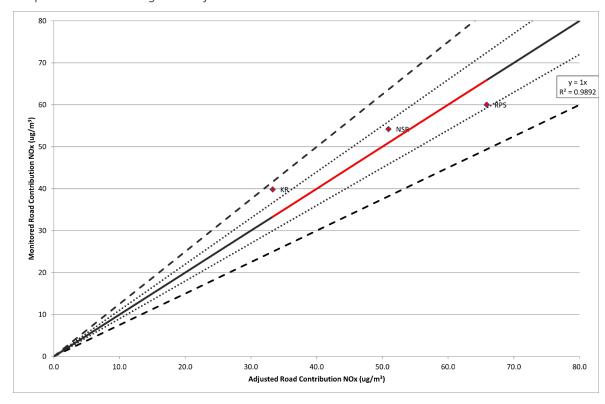
Using the regression correction factor of 1.6551, as identified in Graph 4, it is necessary to adjust the modelled road-NOx. This adjustment factor has been applied to the modelled road-NOx, and the monitored road-NOx has been plotted against adjusted modelled road-NOx (Graph 5). This has then been converted to NO₂, as shown in Table 7 and Graph 6.

Table 7: Verification Procedure: Adjusted Model Results Comparison

| Location | Adjustment applied | 2018 monitored total NO ₂ (μg/m³) | 2018 modelled total NO ₂ (μg/m³) | % Difference (modelled - monitored / monitored) |
|----------|-----------------------|--|--|--|
| KR | 1.6551 | 32 | 28.92 | -9.62% |
| NSR | 1.6551 | 38 | 36.54 | -3.84% |
| RPS | 1.6551 | 42.4 | 44.86 | 5.80% |

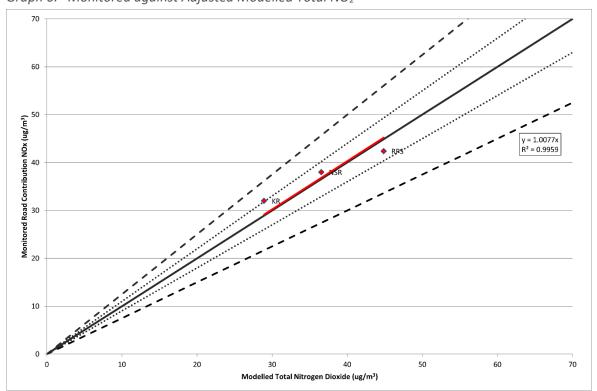
Note:

All NOx to NO_2 conversions undertaken using DEFRA's NOx to NO_2 calculator V8.1, for 2018 emissions and using the 'All other urban UK traffic' traffic mix setting.



Graph 5: Monitored against Adjusted Modelled Road–NOx

Finally, the total monitored NO_2 has been plotted against the adjusted modelled total NO_2 , as presented in Graph 6.



Graph 6: Monitored against Adjusted Modelled Total NO₂

Table 7 and Graph 6 show that the adjusted NO_2 results are within 10% of the monitored NO_2 at all three monitoring sites. Therefore, following adjustment, the model is performing well.



No representative monitoring of PM_{10} or $PM_{2.5}$ is available. To ensure a robust assessment, the adjustment factors calculated for annual mean nitrogen dioxide have also been applied to the modelled concentrations of road PM_{10} and $PM_{2.5}$ at the appropriate receptors in line with guidance set out in LAQM.TG(16).

The supporting documentation for AQC's CREAM V1A explains that the ammonia emissions factors obtained from CREAM V1A will often be used as inputs to ADMS-Roads, but model users will often not be able to verify calculation of ammonia emissions from vehicles due to a lack of roadside ammonia monitoring.

As AQC acknowledge that users will typically not be able to undertake model verification, the documentation includes details of calibration against measurements taken from summer 2014 to summer 2016 at 29 sites in the Ashdown Forest. This shows that the emissions factors obtained using CREAM V1A align well with measurements. This is in contrast to emissions of NOx, which have historically been shown to be under-predicted by DEFRA's EFT. Therefore, it is not considered appropriate to apply the adjustment factor for NOx to emissions of ammonia, as this would likely result in significant over-prediction of ammonia emissions from vehicles.



3 Results

3.1 Human health

Detailed results tables showing the impact at human receptors are provided in Appendix B. This includes the contribution from process emissions associated with the operation of LSEP.

3.1.1 Nitrogen dioxide

As shown in Table 9, for the worst-case scenario, which assumes that the vehicle fleet does not change from the existing mix, the maximum predicted annual mean nitrogen dioxide concentration associated with the traffic from the operation of the LSEP at modelled receptors is $1.81 \, \mu g/m^3$ (or 4.53% of the AQAL) at RR28, along the A530 through Broken Cross, where the majority of LSEP vehicles will pass. When the contribution from process emissions is added to the road contribution the maximum annual mean nitrogen dioxide impact is $1.91 \, \mu g/m^3$ (or 4.53% of the AQAL). RR30, also along this section of the A530, shows similar values.

The predicted environmental concentration (PEC) of nitrogen dioxide emissions at RR28, under the Do Minimum scenario, which incorporates natural growth and other local developments, is 78.6% of the AQAL. The PEC under the Do Something scenario, which includes the LSEP, is 83.3% of the AQAL.

Excluding RR28 and RR30, the PCs at all other receptors are less than 5.5% of the AQAL and PECs are below 75% of the AQAL.

Additional analysis has been carried out for short term impacts. DEFRA's Local Air Quality Management Technical Guidance Note 16 (LAQM.(TG16)) states that if annual mean nitrogen dioxide concentrations are above 60 $\mu g/m^3$ (i.e. 150% of the AQAL), there is the potential for exceedences of the 1-hour AQAL. Even in the worst-case scenario that the fleet mix does not change from current levels the maximum predicted concentration is well below 60 $\mu g/m^3$. Therefore, there is no potential for exceedences of the 1-hour nitrogen dioxide AQAL.

Although the model has been verified to the best of ability providing the local monitoring data and locations, a degree of uncertainty should be given when interpreting the results. However, the magnitude of change results provide some flexibility – there is room for increases in the road concentrations and baseline before impacts would be considered as moderate adverse at the receptors of maximum impact.

During the verification process, it has been noted that the background concentrations from the DEFRA mapped background data are lower than the monitored concentrations. If the actual background concentrations are in fact more in line with the monitored concentrations, this would mean a smaller difference between the roads impacts with LSEP contributions when compared to the baseline. If this were the case, although the background and PECs would be higher, the results of the LSEP impacts would be lower than presented in these results.

Furthermore, the results presented are for the worst-case emissions scenario, in which there is no change to the fleet composition on local networks between 2018 and 2023 and using the 2018 background concentrations. This is a conservative approach as there will be some changeover of the fleet with newer cleaner vehicles and background concentrations are predicted to decrease in future.



3.1.2 Particulate matter

As shown in Table 10, for the worst-case scenario which assumes that the vehicle fleet does not change from the existing mix, the maximum predicted annual mean particulate matter concentration (as PM_{10} or $PM_{2.5}$) associated with the vehicles from LSEP is $0.28~\mu g/m^3$ (0.69% of the AQAL), and $0.17~\mu g/m^3$ (0.83% of the AQAL) respectively. When the contribution from process emissions is added to the road contribution the maximum annual mean particulate matter concentration (as PM_{10} or $PM_{2.5}$) impact increases so minutely and the impact to two decimal places remains at $0.28~\mu g/m^3$ (or 0.0.70% of the AQAL) and $0.17~\mu g/m^3$ (or 0.84% of the AQAL). The maximum PEC for the Do Something scenario at all roads receptors is 36.77% of the AQAL for PM_{10} and 46.81% of the AQAL for $PM_{2.5}$.

As detailed in Appendix 5.2 of the EIA Report, the World Health Organisation (WHO) recommends guidelines for particulate matter which are more stringent than those currently set in UK legislation.

The WHO annual mean PM guidelines values for annual means are as follows:

- Fine particulate matter (PM_{2.5}) 10 μg/m3 as an annual mean.
- Course particulate matter $(PM_{10}) 20 \mu g/m3$ as an annual mean.

The Environment Bill introduces a duty to set a legally binding target for PM_{2.5} although to date this has not been set. For completeness, the results of the maximum of all roads receptors has been compared to the WHO limits guidelines.

Using the WHO limits, the maximum predicted annual mean particulate matter concentration (as PM_{10} or $PM_{2.5}$) associated with the vehicles from LSEP is 1.38% of the AQAL, and 1.67% of the AQAL respectively. When the contribution from process emissions is added to the road contribution the maximum annual mean particulate matter concentration (as PM_{10} or $PM_{2.5}$) impact increases slightly to 1.40% of the AQAL and 1.69% of the AQAL. The maximum PEC for the Do Something scenario at all roads receptors is 68.57% of the AQAL for PM_{10} and 93.61% of the AQAL for $PM_{2.5}$. This conservatively assumes that the entire dust emissions consist of only PM10 or PM2.5s.

3.2 Ecological receptors

There are four ecological sites within 200 m of the roads used by vehicles associated with the LSEP. These are

- Wade Brook LWS;
- Long Wood LWS;
- Rudheath Lime Beds pLWS; and
- Winnington Wood LWS and ancient woodland.

To assess the impacts at these sites, 200 m transects from the edge of the ecological site closest to the road have been modelled, at right angles with the road. The results at each meter along the transect have then been assessed. The transect points have also been entered as receptors into the process emissions from the stack model, to give an overall impact of the LSEP. Due to their size and locations, 3 transects have been produced for Rudheath Lime Beds pLWS and 2 transects for Winnington Woods LWS.

For each transect, the results of the PC and PEC has been presented for annual mean oxides of nitrogen, annual mean ammonia, nitrogen deposition and acid deposition, which are the pollutants from vehicle emissions which have the potential to impact ecological sites. The Do Minimum scenario is based on the predicted 2023 traffic data, incorporating natural growth and new developments and including background levels, and the Do Something scenario is the Do Minimum plus the impact of the LSEP. This is the total number of vehicles associated with the operation of



the LSEP not just the change in vehicles associated with the increase in throughput (i.e. the Proposal).

The verification factor determined for oxides of nitrogen emissions from vehicles (as set out in Section 2.11) has been applied. No additional adjustment has been carried out of the ammonia emissions. The results are based on the worst case scenario that there are no changes to the fleet mix since 2018 – i.e. there is no reduction in emissions in line with projections. As with emissions of NOx, PM₁₀ and PM_{2.5} emissions of ammonia are predicted to change in future years. However, the emissions are not necessarily predicted to decrease in the future and when calculating nitrogen deposition, reductions in NOx may be counteracted by increases in ammonia. As such, a sensitivity analysis has been carried out to determine the difference between using the 2018 and 2023 emission factors for all pollutants. This has shown that the projected reduction in road NOx is offset by an increase in ammonia from vehicles and the nitrogen and acid deposition is similar between the 2018 and 2023 emissions scenarios. The impacts of ammonia are slightly higher using the 2023 rather than the 2018 emission factors. As such all results for NOx, nitrogen and acid deposition are presented using the 2018 emission factors, and ammonia using the 2023 emission factors. Projected ammonia impacts from traffic are not significantly different year on year from 2023. Therefore, it is considered that the use of 2023 emission factors as the maximum impact over the lifetime of the LSEP is appropriate.

When considering the impact of nitrogen and acid deposition the DEFRA NOx to NO2 converter has been used to convert roadside NOx into NO2. This only provides results to 2 decimal places. As such there are small differences between points causing a spiky effect in the graph.

3.2.1 Wade Brook LWS

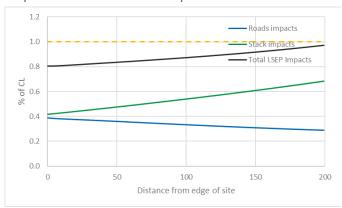
The following section sets out the combined impact of stack emissions and road vehicles along the transect marked in Figure 1 at Wade Brook LWS. Impacts have been presented for the contribution from the roads and process individually and combined, and the PEC which includes the local background concentration. Where appropriate, reference has been made to the 1% and 70% screening criteria.

The impacts have been modelled from the edge of the site, which is 2.9 m from the edge of the road.

Graph 7 to Graph 12 show that the impact of road pollutants decreases with distance away from the road. However, due to the location of Wade Brook LWS in relation to the stack and stack dispersion, the impact of emissions from the LSEP stack, and also overall LSEP impact, increases with distance from the road. This shows that the main contribution to Wade Brook LWS is from the stack rather than the roads. The maximum impacts at Wade Brook LWS are therefore as set out in Appendix 5.2 of the EIA Report (1.5% of the Critical Level for oxides of nitrogen).

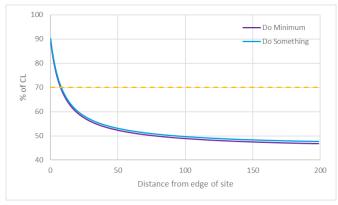
The PECs for ammonia impacts are above the Critical Level for both lichen sensitive communities and non-lichen sensitive communities. However, this is due to high background levels, and the graphs show the minimal difference between Do Minimum impacts, which include general predicted traffic growth and contribution from other developments by 2023, and the Do Something impacts, which includes the contributions from the LSEP (stack and vehicles).

Graph 7: Annual mean NOx impact - PC



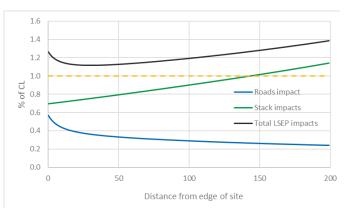
Note: $CL = 30 \mu g/m^3$

Graph 8: Annual mean NOx impact - PEC



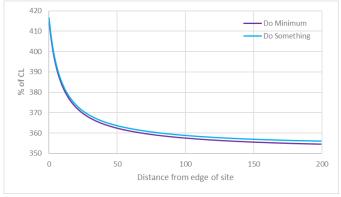
Note: $CL = 30 \mu g/m^3$. $Bg = 11.42 \mu g/m^3$

Graph 9:Annual mean ammonia impact – PC



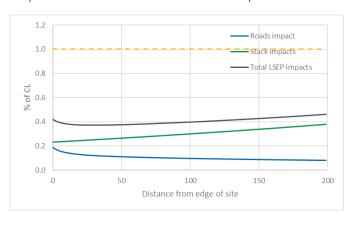
Note: $CL = 1 \mu g/m^{3-}$ for lichen sensitive habitats

Graph 10: Annual mean ammonia impact – PEC



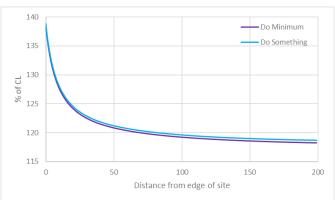
Note: $CL = 1 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 11: Annual mean ammonia impact – PC



Note: $CL = 3\mu g/m^3$ for non lichen sensitive habitats

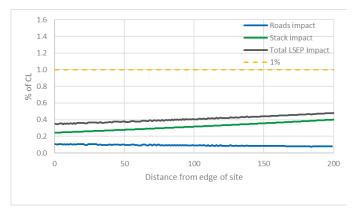
Graph 12: Annual mean ammonia impact – PEC



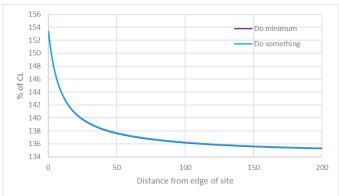
Note: $CL = 3 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 13 to Graph 16 show the deposition impacts. Results for nitrogen deposition have been calculated as a percentage of the lower Critical Load for low and medium altitude hay meadows (10 kgN/ha/yr), noting that the upper Critical Load is 20 kgN/ha/yr. Results for acid deposition have been calculated as a percentage of the lower Critical Load for calcareous grassland (3.017 KeqN/ha/yr).

Graph 13: Annual mean N dep - Grasslands- PC



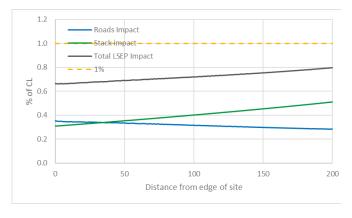
Graph 14: Annual mean N dep –Grasslands PEC



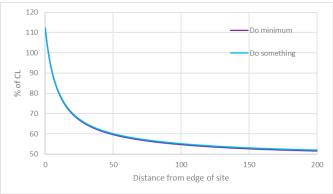
Note:. Lower CL = 20 kgN/ha/yr. Bg = 26.46 kgN/ha/yr

Note: Lower CL = 20 kgN/ha/yr. Bg = 26.46 kgN/ha/yr

Graph 15: Annual mean acid dep - Grasslands - PC



Graph 16: Annual mean acid dep - Grasslands -PEC



Note:. CLmaxN = 5.017 KeqN/ha/yr. Bg: = N 1.89, S 0.21 keq/ha/yr

Note:. $CLmaxN = 5.017 \ KeqN/ha/yr$. $Bg: = N \ 1.89$, $S \ 0.21 \ keq/ha/yr$

For deposition impacts the results show that the impact of road pollutants decreases with distance away from the road, but the stack impact, and also overall LSEP impact, increases with distance from the road. This shows that the main contribution to Wade Brook LWS is from the stack rather than the roads.

Results for deposition show that the impact of the LSEP (roads and stack) is below the 1% screening criteria. The results for the stack emissions show that the maximum impact of stack emissions is 0.9% of the lower Critical Level for nitrogen deposition, and 0.6% of the lower Critical Load for acid deposition. This analysis demonstrates that the inclusion of emissions from traffic associated with the LSEP does not change the conclusions that the impact of nitrogen and acid deposition across the site is less than 1% of the Critical Load



3.2.2 Rudheath Lime Beds pLWS – Transect 1

The following section sets out the combined impact of stack emissions and road vehicles along the transect 1 marked in Figure 1 at Rudheath Lime Beds LWS. Impacts have been presented for the contribution from the roads and process individually and combined, and the PEC which includes the local background concentration. Where appropriate, reference has been made to the 1% and 70% screening criteria.

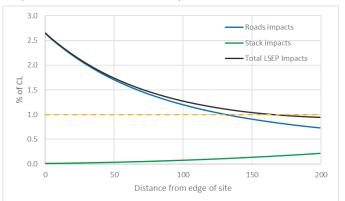
The impacts have been modelled from the edge of the ecological site, which is 2.0 m from the edge of the road.

Graph 17 to Graph 22 show that the impact of road and stack emissions decreases with distance away from the road. The LSEP impact at this location is more influenced by road emissions than stack emissions. The maximum stack contributions at Rudheath Lime Beds are as set out in Appendix 5.2 of the EIA Report (1.3% of the Critical Level for oxides of nitrogen and 2.2% of the Critical Level for annual mean ammonia using the lower Critical Level of 1 μ g/m³). However, this point of maximum impact occurs to the north east edge of the ecological site and is lower at the location of the transect.

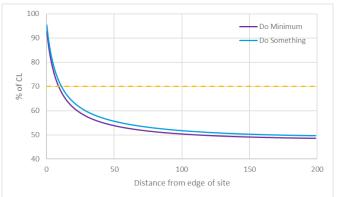
For oxides of nitrogen impacts, although the roads impact does not fall to under 1% of the Critical Level until 133 m from the edge of the ecological site, the PEC of the Do Something scenario falls below 70% of the Critical Level at 11 m from edge of the site. Furthermore, this is only 2 m greater than the distance at which the PEC would fall below 70% of the Critical Level under the Do Minimum scenario.

For ammonia impacts, using the higher Critical Level of $3 \mu g/m^3$, all impacts are below the 1% screening criteria. When using the lower Critical Level of $1 \mu g/m^3$ for lichen sensitive communities, the roads impact falls below the 1% of the Critical Level at 91 m from edge of the ecological site and total impact of the LSEP (roads and stack emissions) fall below 1% of the Critical Level at 129 m from edge of the ecological site. The PEC for ammonia, at either Critical Level, is above the 70% screening criteria due to high background levels. However, Graph 20 and Graph 22 show that impact decreases rapidly with distance from edge of the ecological site and that the difference between the Do Minimum and Do Something scenarios is small.

Graph 17: Annual mean NOx impact - PC



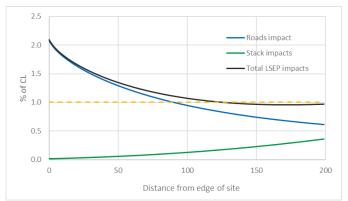
Graph 18: Annual mean NOx impact - PEC



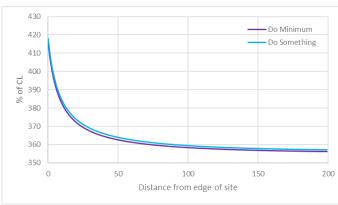
Note: $CL = 30 \mu g/m^3$

Note: $CL = 30 \mu g/m^3$. $Bg = 11.42 \mu g/m^3$

Graph 19: Annual mean ammonia impact – PC



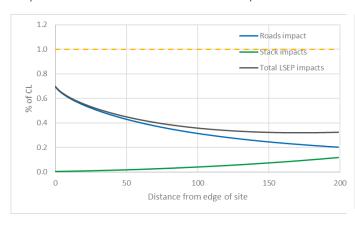
Graph 20: Annual mean ammonia impact – PEC



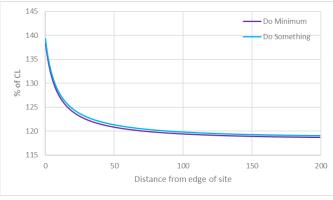
Note: $CL = 1 \mu g/m^{3-}$ for lichen sensitive habitats

Note: $CL = 1 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 21: Annual mean ammonia impact – PC



Graph 22: Annual mean ammonia impact – PEC

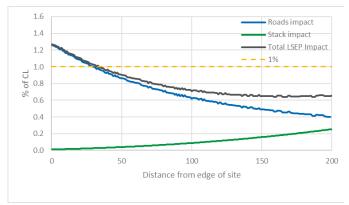


Note: $CL = 3\mu g/m^3$ for non lichen sensitive habitats

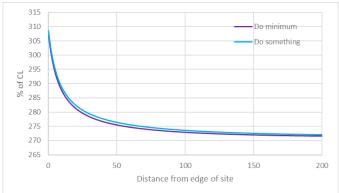
Note: $CL = 3 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 23 to Graph 26 below show the deposition impacts. Results for nitrogen deposition have been calculated as a percentage of the lower Critical Load for valley mires, poor fens and transition mires (10 kgN/ha/yr), noting that the upper Critical Load is 15 kgN/ha/yr. Results for acid deposition have been calculated as a percentage of the lower Critical Load for calcareous grassland (5.017 KeqN/ha/yr).

Graph 23: Annual mean N dep – Grasslands – PC



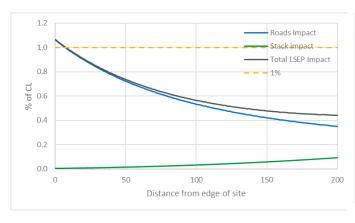
Graph 24: Annual mean N dep – Grasslands – PEC



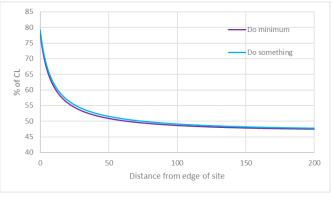
Note:. Lower CL = 10 kgN/ha/yr Bg = 26.46 kgN/ha/yr

Note:. Lower CL = 10 kgN/ha/yr Bg = 26.46 kgN/ha/yr

Graph 25: Annual mean acid dep -Grassland- PC



Graph 26: Annual mean acid dep -Grassland- PEC



Note: CLmaxN = 5.071 KeqN/ha/yr. Bg = N 1.89, S 0.21 keq/ha/yr

Note: CLmaxN = 5.071 KeqN/ha/yr. Bg = N 1.89, S 0.21 keq/ha/yr

For deposition impacts the results show that the impact of road pollutants decreases with distance away from the road and the stack impact increases with distance away from the road. The total LSEP impact is more greatly influenced by the road emissions than the stack emissions for this transect.

For nitrogen deposition, the road impact falls below 1% of the Critical Load at 31 m from the edge of the ecological site, and the total LSEP impact falls below 1% at 34 m from the edge of the ecological site. Nitrogen deposition PEC is well above the 70% screening criteria due to high background levels, however Graph 24 shows that the difference between the Do Minimum and Do Something scenarios is small. For acid deposition, the road impact falls below 1% of the Critical Load at 8 m from the edge of the ecological site and total LSEP impacts fall below 1% of the Critical Load at 9 m from the edge of the ecological site. The PEC for the Do Something scenario falls below the 70% screening criteria at 5 m from the edge of the ecological site.



This analysis demonstrates that the inclusion of emissions from traffic associated with the LSEP means that the impact within 34 m of the edge of the ecological site is greater than 1% of the Critical Load. However, across the rest of the ecological site the impact of nitrogen and acid deposition is less than 1% of the Critical Load.

3.2.3 Rudheath Lime Beds pLWS – Transect 2

The following section sets out the combined impact of process emissions and road vehicles along the transect 2 marked in Figure 1 at Rudheath Lime Beds. The location of this transect is perpendicular to the section of road which the majority of LSEP traffic will be routed. Impacts have been presented for the contribution from the roads and stack emissions individually and combined, and the PEC which includes the local background concentration. Where appropriate, reference has been made to the 1% and 70% screening criteria.

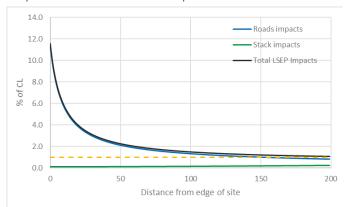
The impacts have been modelled from the edge of the ecological site, which is 2.0 m from the edge of the road.

Graph 27 to Graph 32 show that the impact of road pollutants rapidly decreases with distance away from the road and stack pollutants very slightly increase with distance away from the road. The total LSEP impact at this location is predominantly influenced by road impacts rather than stack impacts. The maximum stack contributions at Rudheath Lime Beds are as set out in Appendix 5.2 of the EIA Report (1.3% of the Critical Level for oxides of nitrogen and 2.2% of the Critical Level for annual mean ammonia using the lower Critical Level of 1 μ g/m³). However, this point of maximum impact occurs to the north east edge of the ecological site and is lower at the location of the transect.

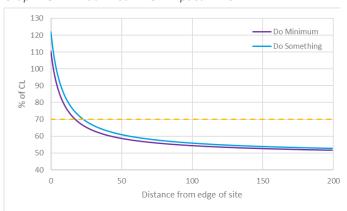
For oxides of nitrogen impacts, although the roads impact does not fall to under 1% of the Critical Level until 153 m from the edge of the ecological site and total LSEP impacts do not fall to under 1% of the Critical Level within the 200 m transect, the PEC of the Do Something scenario falls below 70% of the Critical Level at 24 m from edge of the site. This is 6 m greater than the distance at which the PEC would fall below 70% of the Critical Level under the Do Minimum scenario.

For ammonia impacts, using the higher Critical Level of $3 \mu g/m^3$, road impacts fall below the 1% screening criteria at 43 m from the edge of the ecological site and total LSEP impacts fall below the 1% at 48 m from the edge of the ecological site. When using the lower Critical Level of $1 \mu g/m^3$ for lichen sensitive communities, the roads impact does not fall below the 1% of the Critical Level until 178 m from edge of the ecological site and total LSEP impacts do not fall below 1% of the Critical Level within the 200 m transect. The PEC for ammonia, at either Critical Level, is above the 70% screening criteria due to high background levels. However, Graph 30 and Graph 32 show that impact decreases rapidly with distance from edge of the ecological site and that the difference between the Do Minimum and Do Something scenarios is slight.

Graph 27: Annual mean NOx impact - PC

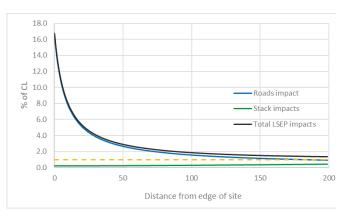


Graph 28: Annual mean NOx impact - PEC



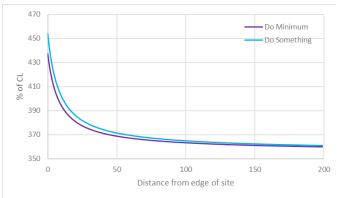
Note: $CL = 30 \mu g/m^3$

Graph 29: Annual mean ammonia impact – PC



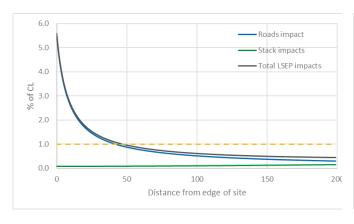
Note: $CL = 30 \mu g/m^3$. $Bg = 11.42 \mu g/m^3$

Graph 30: Annual mean ammonia impact – PEC



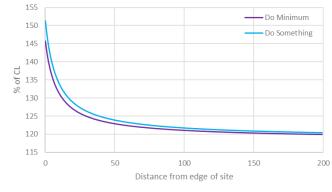
Note: $CL = 1 \mu g/m^{3-}$ for lichen sensitive habitats

Graph 31: Annual mean ammonia impact – PC



Note: $CL = 1 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 32: Annual mean ammonia impact – PEC

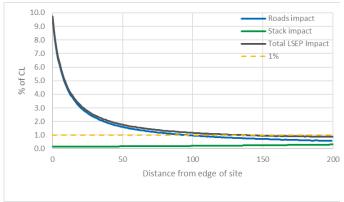


Note: $CL = 3\mu g/m^3$ for non lichen sensitive habitats

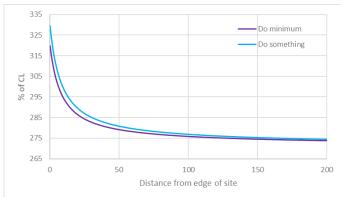
Note: $CL = 3 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 33 to Graph 36 below show the deposition impacts. Results for nitrogen deposition have been calculated as a percentage of the lower Critical Load for valley mires, poor fens and transition mires (10 kgN/ha/yr), noting that the upper Critical Load is 15 kgN/ha/yr. Results for acid deposition have been calculated as a percentage of the lower Critical Load for calcareous grassland (5.017 KeqN/ha/yr).

Graph 33: Annual mean N Dep Grassland- PC



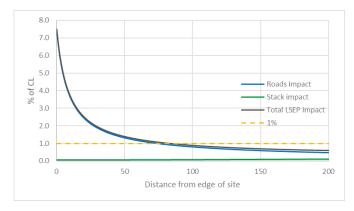
Graph 34: Annual mean N Dep Grassland- PEC



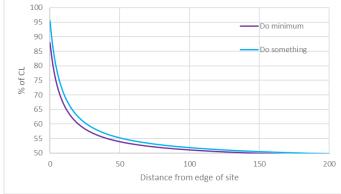
Note:. Lower CL = 10 kgN/ha/yr Bg = 26.46 kgN/ha/yr

Note:. Lower CL = 10 kgN/ha/yr Bg = 26.46 kgN/ha/yr

Graph 35: Annual mean acid dep - Grassland- PC



Graph 36: Annual mean acid dep - Grassland- PEC



Note:. CLmaxN = 5.071 KeqN/ha/yr, Bg = N 1.89, S 0.21 keq/ha/yr

Note:. CLmaxN = 5.071 KeqN/ha/yr, Bg = N 1.89, S 0.21 keq/ha/yr

For deposition impacts the results show that the impact of road pollutants decreases rapidly with distance away from the road. The stack impact at this location is not very influential on the total LSEP impact.

For nitrogen deposition, the road impact falls below 1% of the Critical Load at 96 m from the edge of the ecological site, and the total LSEP impact falls below 1% at 136 m from the edge of the ecological site. Nitrogen deposition PEC is well above the 70% screening criteria due to high background levels, however Graph 34 shows that the difference between the Do Minimum and Do Something scenarios is small. For acid deposition, the road impact falls below 1% of the Critical Load at 76 m from the edge of the ecological site, and the total LSEP impact falls below 1% at 84 m from the edge of the ecological site. The PEC for the Do Something scenario rapidly decreases with distance from the road and falls below the 70% screening criteria at 11 m from the edge of the ecological site.



This analysis demonstrates that the inclusion of emissions from traffic associated with the LSEP is the main contribution to the impact on Rudheath Lime Beds along this section of the A530. This is not unexpected, as the vast majority of all LSEP traffic and all LSEP HGV traffic will be routed down this section of road. As shown above, the impacts drop of rapidly from the edge of the site, and the difference between the Do Minimum and Do Something scenarios shall be assessed for significance within the ecological interpretation (Appendix 5.5).

3.2.4 Rudheath Lime Beds pLWS – Transect 3

The following section sets out the combined impact of process emissions and road vehicles along the transect 3 marked in Figure 1 at Rudheath Lime Beds. The location of this transect perpendicular to the A556, but is set back from the edge of the road by 81.5 m, beginning at the edge of the ecological site. The transect travels from east to west, away from the A556, but towards the LSEP and Griffiths Road. Impacts have been presented for the contribution from the roads and stack emissions individually and combined, and the PEC which includes the local background concentration. Where appropriate, reference has been made to the 1% and 70% screening criteria.

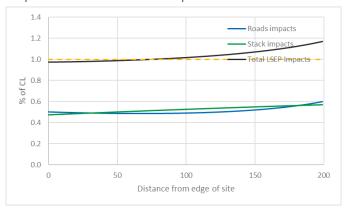
The impacts have been modelled from the edge of the ecological site, which is 81.5 m from the edge of the road.

Graph 37 to Graph 42 show that the impact of road pollutants slightly increases with distance away from the A556. There is not a rapid decrease due to the distance of the ecological site from the edge of the road, and the increase with distance from the A556 is in keeping with the transect getting closer to Griffiths Road, which has the highest vehicle impact from the LSEP. The impact at this transect is influenced by both traffic and the stack emissions at a similar level. The maximum stack contributions at Rudheath Lime Beds are as set out in Appendix 5.2 of the EIA Report (1.3% of the Critical Level for oxides of nitrogen and 2.2% of the Critical Level for annual mean ammonia using the lower Critical Level of $1 \,\mu\text{g/m}^3$). However, this point of maximum impact occurs to the north east edge of the ecological site and is lower at the location of the transect.

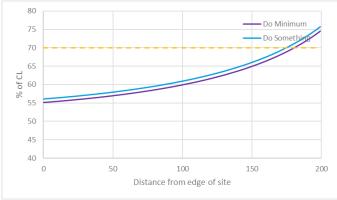
For oxides of nitrogen impacts, both the roads and stack emissions are well below the 1% screening criteria, but when in combination, they exceed the 1% screening criteria closer to the LSEP site. The PEC is below 70% at the western side of the ecological site, but the transect shows it exceeds 70% moving into the centre of the ecological site towards the LSEP.

For ammonia impacts, using the lower Critical Level of $1 \mu g/m^3$ for lichen sensitive communities, roads impacts and stack impacts are below the 1% screening criteria, however, when in combination, exceed the 1% screening criteria. When using the higher Critical Level of $3 \mu g/m^3$, all impacts are below the 1% screening criteria. Due to high background levels, the ammonia PEC is above 70% for both lichen sensitive communities and non-lichen sensitive communities. Graph 40 and Graph 42 show there is a small increase in the impact with the Do Something scenario compared to the Do Minimum scenario.

Graph 37: Annual mean NOx impact - PC



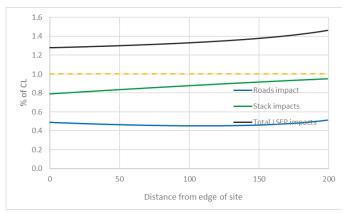
Graph 38: Annual mean NOx impact - PEC



Note: $CL = 30 \mu g/m^3$

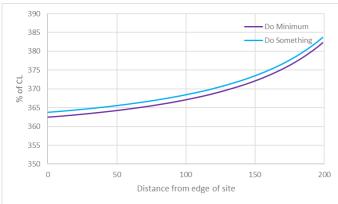
Note: $CL = 30 \mu g/m^3$. $Bg = 11.42 \mu g/m^3$

Graph 39: Annual mean ammonia impact – PC



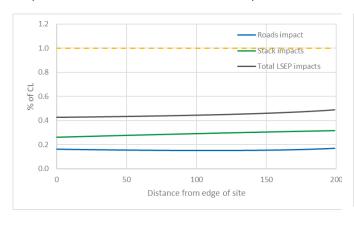
Note: $CL = 1 \mu g/m^{3-}$ for lichen sensitive habitats

Graph 40: Annual mean ammonia impact – PEC



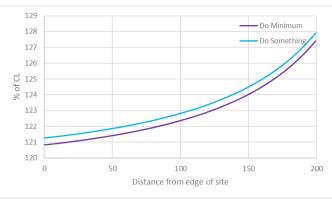
Note: $CL = 1 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 41: Annual mean ammonia impact – PC



Note: $CL = 3\mu g/m^3$ for non lichen sensitive habitats

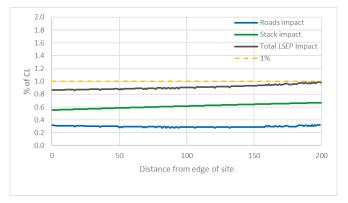
Graph 42: Annual mean ammonia impact – PEC



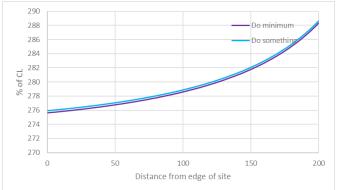
Note: $CL = 3 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 43 to Graph 46 below show the deposition impacts. Results for nitrogen deposition have been calculated as a percentage of the lower Critical Load for valley mires, poor fens and transition mires (10 kgN/ha/yr), noting that the upper Critical Load is 15 kgN/ha/yr. Results for acid deposition have been calculated as a percentage of the lower Critical Load for calcareous grassland (5.017 KeqN/ha/yr).

Graph 43: Annual mean N dep- Grassland- PC



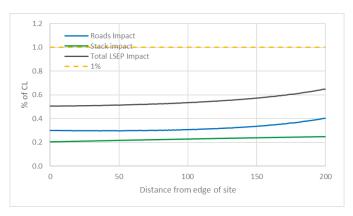
Graph 44: Annual mean N dep- Grassland- PEC



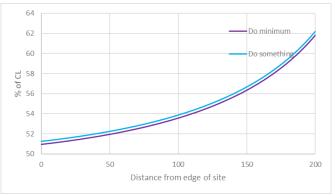
Note:. Lower CL = 10 kgN/ha/yr Bg = 26.46 kgN/ha/yr

Note:. Lower CL = 10 kgN/ha/yr Bg = 26.46 kgN/ha/yr

Graph 45: Annual mean acid dep - Grassland- PC



Graph 46: Annual mean acid dep - Grassland- PEC



Note:. CLmaxN = 5.071 KeqN/ha/yr, Bg = N 1.89, S 0.21 keq/ha/yr

Note:. $CLmaxN = 5.071 \ KeqN/ha/yr, Bg = N \ 1.89, S \ 0.21 \ keq/ha/yr$

For deposition impacts the results show that the impact of both road pollutants and stack pollutants increase very slightly towards the centre of the ecological site. All impacts remain under the 1% screening criteria, although the nitrogen deposition graph is suggestive that the impacts would exceed the 1% if the transect was continued towards the LSEP and Griffiths Road. PEC for nitrogen deposition is over the 70% screening criteria due to high background levels.

This analysis demonstrates that the inclusion of emissions from traffic associated with the LSEP has a contribution to the impact on Rudheath Lime Beds, but suggests that the larger contribution is from the A530 rather than the A556. The vast majority of all LSEP LDV traffic and all LSEP HGV traffic will be routed down this section of the A530.



3.2.5 Long Wood LWS

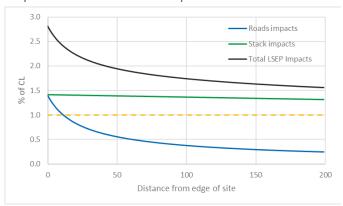
The following section sets out the combined impact of process emissions and road vehicles along the transect marked in Figure 1 at Long Wood. The transect runs from west to east on the eastern side of the A556 near the junction at Lostock Gralam. The transect is not fully perpendicular to the road due to the shape of the ecological site which it fits within. Impacts have been presented for the contribution from the roads and process individually and combined, and the PEC which includes the local background concentration. Where appropriate, reference has been made to the 1% and 70% screening criteria.

The impacts have been modelled from the edge of the ecological site, which is 5.5 m from the edge of the road.

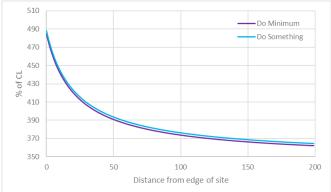
Graph 47 to Graph 52 show that the impact of road pollutants decreases with distance away from the road. Stack pollutant impacts also decrease with distance away from the road, but remain an influence to the in combination impacts. The maximum stack contributions at Long Wood are as set out in Appendix 5.2 of the EIA Report (1.6% of the Critical Level for oxides of nitrogen and 2.6% of the Critical Level for annual mean ammonia using the lower Critical Level of $1 \mu g/m^3$).

For oxides of nitrogen impacts, at the edge of the site, the roads impact is above the 1% screening criteria, but quickly decreases with distance and falls below the 1% screening criteria at 12 m from the edge of the ecological site. However, due to the impact of stack emissions at this location, the in-combination impact is above the 1% screening criteria along the whole transect. A similar pattern is shown for ammonia impacts when using the lower Critical Level of 1 $\mu g/m^3$. When using the higher Critical Level of 3 $\mu g/m^3$, the total LSEP impact exceeds the 1% screening criteria up to 20 m from the edge of the ecological site. The PEC graphs show that although there are exceedances of the 70% screening criteria, there is minimal difference between the Do Minimum and Do Something scenarios. The drop off from the road is more pronounced within the PEC graphs due to the inclusion of other regular traffic.

Graph 47: Annual mean NOx impact - PC

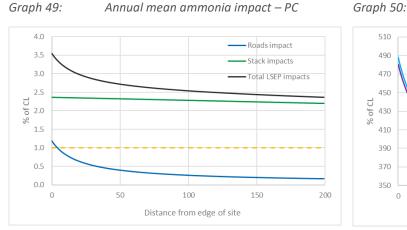


Graph 48: Annual mean NOx impact - PEC

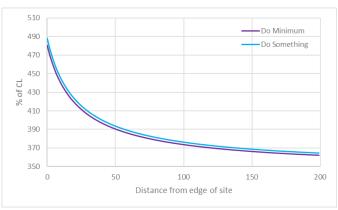


Note: $CL = 30 \mu g/m^3$

Graph 49: Annual mean ammonia impact – PC



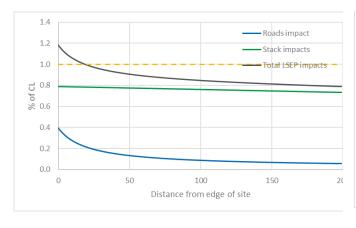
Note: $CL = 30 \mu g/m^3$. $Bg = 11.42 \mu g/m^3$



Annual mean ammonia impact – PEC

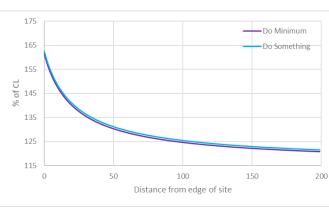
Note: $CL = 1 \mu g/m^{3-}$ for lichen sensitive habitats

Graph 51: Annual mean ammonia impact – PC



Note: $CL = 1 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 52:



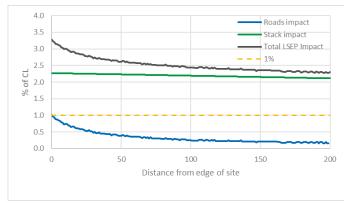
Annual mean ammonia impact – PEC

Note: $CL = 3\mu g/m^3$ for non lichen sensitive habitats

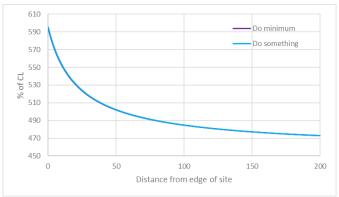
Note: $CL = 3 \mu g/m^3$. $Bg = 3.45 \mu g/m^3$

Graph 53 to Graph 56 below show the deposition impacts. Results for nitrogen deposition have been calculated as a percentage of the lower Critical Load for broadleaved deciduous woodland (10 kgN/ha/yr), noting that the upper Critical Load is 20 kgN/ha/yr. Results for acid deposition have been calculated as a percentage of the lower Critical Load for unmanaged broadleaved coniferous woodland (1.897 KeqN/ha/yr).

Graph 53: Annual mean N dep - Woodlands- PC



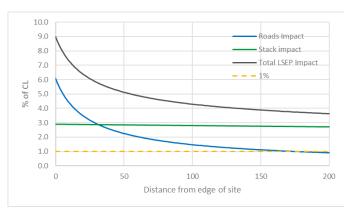
Graph 54: Annual mean N dep - Woodlands- PEC



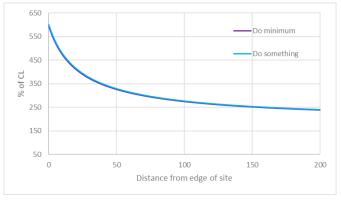
Note:. Lower CL = 10 kgN/ha/yr Bg:= 45.50 kgN/ha/yr

Note:. Lower CL = 10 kgN/ha/yr Bg = 45.50 kgN/ha/yr

Graph 55: Annual mean acid dep Woodland-PC



Graph 56: Annual mean acid dep Woodland-PEC



Note: CLmaxN = 1.897 KeqN/ha/yr, Bg = N 3.25, S 0.25 keq/ha/yr

Note: CLmaxN = 1.897 KeqN/ha/yr, Bg = N 3.25, S 0.25 keq/ha/yr

For deposition impacts the results show that the impact of road pollutants decreases with distance away from the road. Stack pollutant impacts also decrease with distance away from the road, but remain an influence to the in combination impacts. The drop off from the road is more pronounced within the PEC graphs due to the inclusion of other regular traffic.

For nitrogen deposition, although the roads impact is below the 1% screening criteria, the contribution from the stack emissions causes in-combination effects to exceed the criteria.

For acid deposition, the road contribution is above the 1% screening criteria until 179 m from the edge of the ecological site. The PEC for all deposition impacts is well above the 70% screening criteria due to high background levels.

The results for the stack emissions show that the maximum impact of stack emissions is 3.0% of the lower Critical Level for nitrogen deposition, and 3.6% of the lower Critical Load for acid deposition.



This analysis demonstrates that the inclusion of emissions from traffic associated with the LSEP has some contribution to the impact on Long Wood, but suggests that at a distance from the road, the main contribution at this location is from the stack.

3.2.6 Winnington Woods LWS - Transect 1

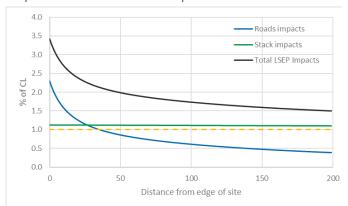
The following section sets out the combined impact of process emissions and road vehicles along the transect 1 marked in Figure 1 at Winnington Woods. The transect runs from south to north perpendicular to the A556. Impacts have been presented for the contribution from the roads and process individually and combined, and the PEC which includes the local background concentration. Where appropriate, reference has been made to the 1% and 70% screening criteria.

The impacts have been modelled from the edge of the ecological site, which is 5.5 m from the edge of the road.

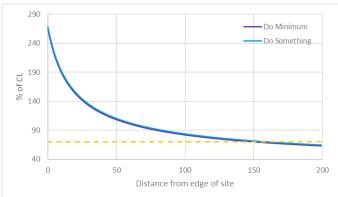
Graph 57 to Graph 62 show that the impact of road pollutants decreases with distance away from the road. At this location and transect orientation in relation to the LSEP, the stack contributions are quite consistent across the transect. The maximum stack contributions at Winnington Wood are as set out in Appendix 5.2 of the EIA Report (1.5% of the Critical Level for oxides of nitrogen and 2.45% of the Critical Level for annual mean ammonia using the lower Critical Level of 1 μ g/m³). These values already exceed the 1% screening criteria, so the in-combination impact is also in exceedance. However, the impact from the road drops off to below the 1% screening criteria at 36 m from the edge of the ecological site for oxides of nitrogen and 20 m from the edge of the ecological site for ammonia (when using the lower Critical Level of 1 μ g/m³). When using the higher Critical Level of 3 μ g/m³, the roads impact is well below the 1% screening criteria and the total LSEP impact falls below the 1% screening criteria at 16 m from the edge of the ecological site. The contributions from the stack at the point of maximum impact at Winnington Wood are as set out in Appendix 5.2 of the EIA Report (1.5% of the Critical Level for oxides of nitrogen and 2.45% for annual mean ammonia using the lower Critical Level of 1 μ g/m³).

The PEC graphs show an exceedences of the 70% screening criteria due to high background levels, but show the minimal difference between the Do Minimum and Do Something scenarios.

Graph 57: Annual mean NOx impact - PC

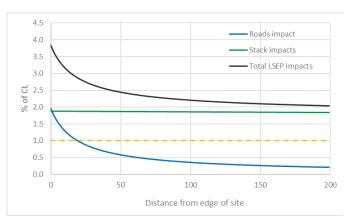


Graph 58: Annual mean NOx impact - PEC



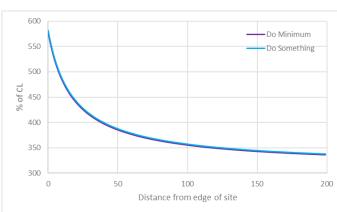
Note: $CL = 30 \mu g/m^3$

Graph 59: Annual mean ammonia impact – PC



Note: $CL = 30 \mu g/m^3$. $Bg = 11.42 \mu g/m^3$

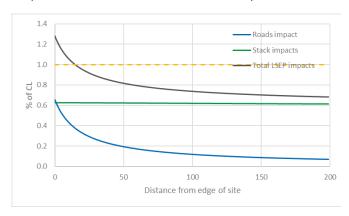
Graph 60:



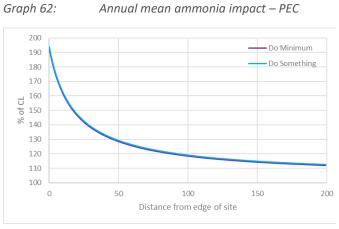
Annual mean ammonia impact – PEC

Note: $CL = 1 \mu g/m^{3-}$ for lichen sensitive habitats

Graph 61: Annual mean ammonia impact – PC



Note: $CL = 1 \mu g/m^3$. $Bg = 3.11 \mu g/m^3$



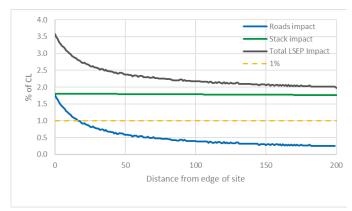
Annual mean ammonia impact – PEC

Note: $CL = 3\mu g/m^3$ for non lichen sensitive habitats

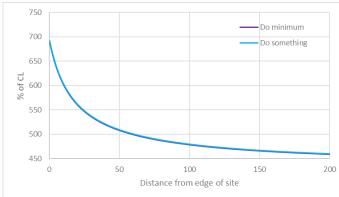
Note: $CL = 3 \mu g/m^3$. $Bg = 3.11 \mu g/m^3$

Graph 63 to Graph 66 below show the deposition impacts. Results for nitrogen deposition have been calculated as a percentage of the lower Critical Load for broadleaved deciduous woodland (10 kgN/ha/yr), noting that the upper Critical Load is 20 kgN/ha/yr. Results for acid deposition have been calculated as a percentage of the lower Critical Load for unmanaged broadleaved coniferous woodland (1.871 KeqN/ha/yr).

Graph 63: Annual mean N dep – Woodlands - PC



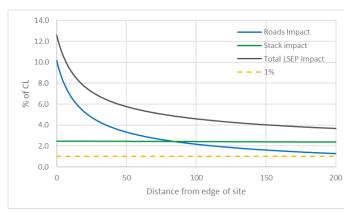
Graph 64: Annual mean N dep – Woodlands - PEC



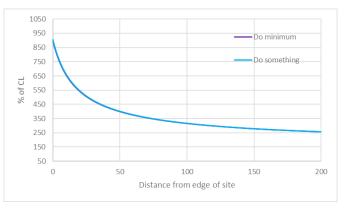
Note: Lower CL = 10 kgN/ha/yr, Bg = 43.40 kgN/ha/yr

Note: Lower CL = 10 kgN/ha/yr, Bg = 43.40 kgN/ha/yr

Graph 65: Annual mean acid dep Woodland- PC



Graph 66: Annual mean acid dep Woodland-PEC



Note: CLmaxN = 1.871 KeqN/ha/yr, Bg = N 3.10, S 0.26 keq/ha/yr

Note: CLmaxN = 1.871 KeqN/ha/yr, Bg = N 3.10, S 0.26 keq/ha/yr

For deposition impacts the results show that the impact of road pollutants decreases with distance away from the road. At this location and transect orientation in relation to the LSEP, the stack contributions are quite consistent across the transect, but have an influence on the in combination impacts. The drop off from the road is more pronounced within the PEC graphs due to the inclusion of other regular traffic.

For nitrogen deposition, although the roads impact falls below the 1% screening criteria at 17 m from the edge of the ecological site, the contribution from the stack emissions causes incombination effects to exceed the 1% screening criteria across the transect. The stack contribution at the point of maximum impact across Winnington Wood is 2.8% of the lower Critical Load.

For acid deposition, the road contribution, and also total LSEP contribution, is above the 1% screening criteria for the full length of the transect. The stack contribution at the point of maximum



impact across Winnington Wood is 3.8% of the lower Critical Load. The PEC for all deposition impacts is well above the 70% screening criteria due to high background levels.

This analysis demonstrates that the inclusion of emissions from traffic associated with the LSEP has some contribution to the impact on Winnington Long Wood, but suggests that at distance from the road, the main constant contribution at this location is from the stack.

3.2.7 Winnington Wood LWS – Transect 2

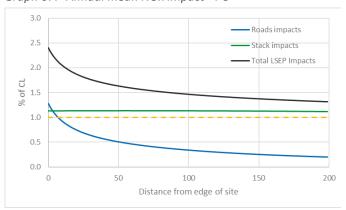
The following section sets out the combined impact of process emissions and road vehicles along the transect 2 marked in Figure 1 at Winnington Woods. The transect runs from north to south, perpendicular to the A556 and opposite Winnington Woods transect 1. Impacts have been presented for the contribution from the roads and process individually and combined, and the PEC which includes the local background concentration. Where appropriate, reference has been made to the 1% and 70% screening criteria.

The impacts have been modelled from the edge of the ecological site, which is 5.5 m from the edge of the road.

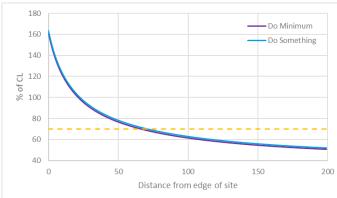
Graph 67 to Graph 72 show that the impact of road pollutants decreases with distance away from the road. At this location and transect orientation in relation to the LSEP, the stack contributions are quite consistent across the transect. The maximum stack contributions at Winnington Wood are as set out in Appendix 5.2 of the EIA Report (1.5% of the Critical Level for oxides of nitrogen and 2.45% of the Critical Level for annual mean ammonia using the lower Critical Level of 1 μ g/m³. These values already exceed the 1% screening criteria, so the in-combination impact is also in exceedance. However, the impact from the road drops off to below the 1% screening criteria at 7 m from the edge of the ecological site for oxides of nitrogen, at 4 m from the edge pf the ecological site for annual mean ammonia (using the lower Critical Level of 1 μ g/m³) and is below the 1% screening criteria for ammonia when using the higher Critical Level of 3 μ g/m³.

The PEC graphs show an exceedences of the 70% screening criteria due to high background levels, but show the minimal difference between the Do Minimum and Do Something scenarios.

Graph 67: Annual mean NOx impact - PC



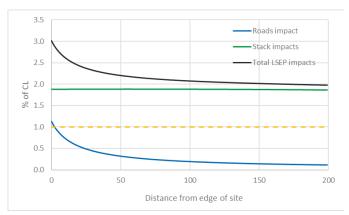
Graph 68: Annual mean NOx impact - PEC



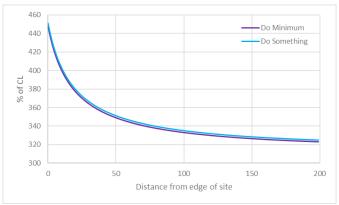
Note: $CL = 30 \mu g/m^3$

Note: $CL = 30 \mu g/m^3$. $Bg = 11.42 \mu g/m^3$

Graph 69: Annual mean ammonia impact – PC



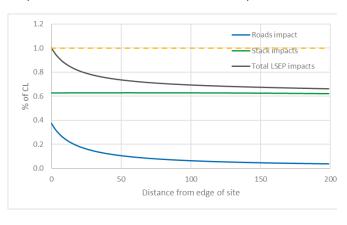
Graph 70: Annual mean ammonia impact – PEC



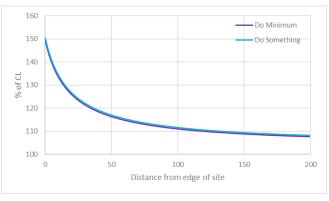
Note: $CL = 1 \mu g/m^{3-}$ for lichen sensitive habitats

Note: $CL = 1 \mu g/m^3$. $Bg = 3.11 \mu g/m^3$

Graph 71: Annual mean ammonia impact – PC



Graph 72: Annual mean ammonia impact – PEC

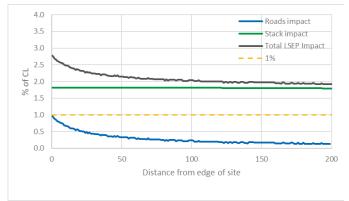


Note: $CL = 3\mu g/m^3$ for non lichen sensitive habitats

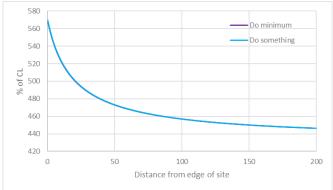
Note: $CL = 3 \mu g/m^3$. $Bg = 3.11 \mu g/m^3$

Graph 73 to Graph 76 below show the deposition impacts. Results for nitrogen deposition have been calculated as a percentage of the lower Critical Load for broadleaved deciduous woodland (10 kgN/ha/yr), noting that the upper Critical Load is 20 kgN/ha/yr. Results for acid deposition have been calculated as a percentage of the lower Critical Load for unmanaged broadleaved coniferous woodland (1.871 KeqN/ha/yr).

Graph 73: Annual mean N dep – Woodlands - PC

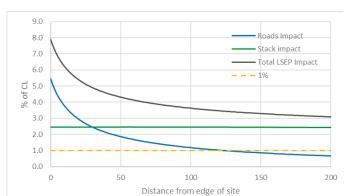


Graph 74: Annual mean N dep – Woodlands - PEC

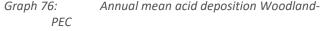


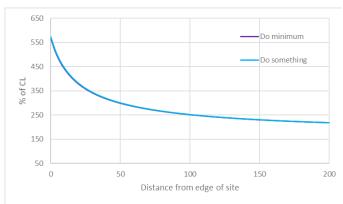
Note: Lower CL = 10 kgN/ha/yr, Bg = 43.40 kgN/ha/yr

Graph 75: Annual mean acid dep Woodland- PC



Note: Annual mean acid dep Woodland-PEC





Note: CLmaxN = 1.871 KeqN/ha/yr, Bg = N 3.10, S 0.26 keq/ha/yr

Note: CLmaxN = 1.871 KeqN/ha/yr, Bg = N 3.10, S 0.26 keq/ha/yr

For deposition impacts the results show that the impact of road pollutants decreases with distance away from the road. At this location and transect orientation in relation to the LSEP facility, the stack contributions are quite consistent across the transect, but an influence to the in combination impacts. The drop off from the road is more pronounced within the PEC graphs due to the inclusion of other regular traffic.

For nitrogen deposition, although the roads impact is below the 1% screening criteria the contribution from the stack emissions causes in-combination effects to exceed the 1% screening criteria across the transect. The stack contribution at the point of maximum impact across Winnington Wood is 2.8% of the lower Critical Load.



For acid deposition, the road contribution falls below the 1% screening criteria at 123 m from the edge of the ecological site, but the total LSEP contribution is above the 1% screening criteria for the full length of the transect. The stack contribution at the point of maximum impact across Winnington Wood is 3.8% of the lower Critical Load. The PEC for all deposition impacts is well above the 70% screening criteria due to high background levels.

This analysis demonstrates that the inclusion of emissions from traffic associated with the LSEP has some contribution to the impact on Winnington Wood, but suggests that at distance from the road, the main contribution at this location is from the stack.



| Annexes | | |
|----------|--|--|
| ABBOVOC | | |
| AMPXES | | |
| MIIICACS | | |
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A Traffic data

Table 8: Traffic Data for Road Links (24-hour AADT)

| Road link | Road name/description | Ва | seline 2018 | Do Mi | nimum 2023 | Do Son | nething 2023 | Develo | pment trips |
|-----------|--|--------|-------------|--------|------------|--------|--------------|--------|-------------|
| | | LDVs | HDVs | LDVs | HDVs | LDVs | HDVs | LDVs | HDVs |
| Α | Site Access Road | 1,289 | 277 | 1,593 | 268 | 1,674 | 702 | 82 | 434 |
| В | Griffiths Road, north of Site entrance | 7,722 | 0 | 8,093 | 0 | 8,107 | 0 | 14 | 0 |
| С | Griffiths Road, south of Site entrance | 8,021 | 277 | 8,363 | 268 | 8,430 | 702 | 67 | 434 |
| D | Middlewich Road into Northwich | 9,692 | 265 | 9,650 | 244 | 9,650 | 244 | 0 | 0 |
| Е | Penny Lane | 0 | 0 | 0 | 0 | 0 | 0 | 0 | 0 |
| F | A350 south of Middlewich Road Junction | 15,616 | 535 | 15,874 | 502 | 15,942 | 935 | 67 | 434 |
| Ga | A556, west of roundabout eastbound | 13,341 | 669 | 13,454 | 630 | 13,455 | 706 | 1 | 76 |
| Gb | A556, west of roundabout westbound | 17,214 | 686 | 16,328 | 657 | 16,329 | 733 | 1 | 76 |
| На | A350 south of roundabout, northbound | 6,907 | 315 | 6,733 | 297 | 6,755 | 362 | 22 | 65 |
| Hb | A350 south of roundabout, southbound | 6,261 | 241 | 6,527 | 229 | 6,549 | 294 | 22 | 65 |
| Н | A350 south of roundabout, single carriageway | 13,168 | 556 | 13,260 | 526 | 13,305 | 656 | 44 | 130 |
| la | A556, east of roundabout eastbound | 15,828 | 881 | 14,240 | 744 | 14,250 | 820 | 10 | 76 |
| Ib | A556, east of roundabout westbound | 17,377 | 1,101 | 14,925 | 926 | 14,935 | 1002 | 10 | 76 |

Note:

Figures subject to minor rounding errors

Development trips equates to the total throughput of the LSEP rather than just the increase.

The Do-Minimum scenario excludes the contribution associated with the section 36 consent.



B Vehicle emissions detailed results tables- human receptors

Table 9: Annual Mean Nitrogen Dioxide – Worst Case

| Receptor | | Do-Minimum | | LSEP | | Do-Something | | | Impact |
|----------|-------|------------|----------------------------|-------|-------|--------------|-------|--------|--------------------------------------|
| | | | Stack PC Road contribution | | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| RR1 | 19.18 | 48.0% | 0.28 | 0.10 | 19.56 | 48.91% | 0.38 | 0.96% | negligible |
| RR2 | 18.35 | 45.9% | 0.30 | 0.08 | 18.73 | 46.83% | 0.38 | 0.95% | negligible |
| RR3 | 18.32 | 45.8% | 0.31 | 0.08 | 18.71 | 46.77% | 0.39 | 0.97% | negligible |
| RR4 | 18.29 | 45.7% | 0.31 | 0.08 | 18.68 | 46.70% | 0.39 | 0.97% | negligible |
| RR5 | 21.04 | 52.6% | 0.31 | 0.12 | 21.47 | 53.68% | 0.43 | 1.08% | negligible |
| RR6 | 13.61 | 34.0% | 0.32 | 0.04 | 13.97 | 34.93% | 0.36 | 0.91% | negligible |
| RR7 | 26.52 | 66.3% | 0.37 | 0.19 | 27.08 | 67.70% | 0.56 | 1.40% | negligible |
| RR8 | 26.91 | 67.3% | 0.37 | 0.22 | 27.50 | 68.75% | 0.59 | 1.48% | negligible |
| RR9 | 23.26 | 58.2% | 0.36 | 0.15 | 23.77 | 59.44% | 0.51 | 1.29% | negligible |
| RR10 | 23.9 | 59.8% | 0.35 | 0.16 | 24.41 | 61.02% | 0.51 | 1.27% | negligible |
| RR11 | 24.3 | 60.8% | 0.34 | 0.17 | 24.81 | 62.03% | 0.51 | 1.28% | negligible |
| RR12 | 18.12 | 45.3% | 0.19 | 0.10 | 18.41 | 46.02% | 0.29 | 0.72% | negligible |
| RR13 | 18.38 | 46.0% | 0.17 | 0.11 | 18.66 | 46.64% | 0.28 | 0.69% | negligible |
| RR14 | 14.39 | 36.0% | 0.06 | 0.06 | 14.51 | 36.28% | 0.12 | 0.30% | negligible* |
| RR15 | 25.02 | 62.6% | 0.12 | 0.14 | 25.28 | 63.21% | 0.26 | 0.66% | negligible |
| RR16 | 27.3 | 68.3% | 0.13 | 0.31 | 27.74 | 69.34% | 0.44 | 1.09% | negligible |



| Receptor | 1 | Do-Minimum | | LSEP | | Do-Something | | | Impact |
|----------|-------|------------|----------|-------------------|-------|--------------|-------|--------|--------------------------------|
| | | | Stack PC | Road contribution | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| RR17 | 28.57 | 71.4% | 0.12 | 0.37 | 29.06 | 72.66% | 0.49 | 1.23% | negligible |
| RR18 | 29.35 | 73.4% | 0.12 | 0.41 | 29.88 | 74.70% | 0.53 | 1.32% | negligible |
| RR19 | 25.37 | 63.4% | 0.06 | 0.25 | 25.68 | 64.19% | 0.31 | 0.77% | negligible |
| RR20 | 24.34 | 60.9% | 0.06 | 0.28 | 24.68 | 61.69% | 0.34 | 0.84% | negligible |
| RR21 | 23.81 | 59.5% | 0.07 | 0.19 | 24.07 | 60.18% | 0.26 | 0.65% | negligible |
| RR22 | 23.56 | 58.9% | 0.07 | 0.22 | 23.85 | 59.62% | 0.29 | 0.72% | negligible |
| RR23 | 25.11 | 62.8% | 0.06 | 0.28 | 25.45 | 63.64% | 0.34 | 0.86% | negligible |
| RR24 | 27.58 | 69.0% | 0.06 | 0.74 | 28.38 | 70.95% | 0.80 | 2.00% | negligible |
| RR25 | 30.11 | 75.3% | 0.06 | 0.46 | 30.63 | 76.57% | 0.52 | 1.30% | negligible |
| RR26 | 26.62 | 66.6% | 0.06 | 0.85 | 27.53 | 68.83% | 0.91 | 2.28% | negligible |
| RR27 | 22.92 | 57.3% | 0.06 | 0.66 | 23.64 | 59.10% | 0.72 | 1.80% | negligible |
| RR28 | 31.43 | 78.6% | 0.06 | 1.81 | 33.30 | 83.25% | 1.87 | 4.68% | slight, adverse |
| RR29 | 24.41 | 61.0% | 0.06 | 1.09 | 25.56 | 63.90% | 1.15 | 2.88% | negligible |
| RR30 | 30.38 | 76.0% | 0.06 | 1.74 | 32.18 | 80.45% | 1.80 | 4.50% | slight, adverse |
| RR31 | 21.95 | 54.9% | 0.06 | 0.92 | 22.93 | 57.33% | 0.98 | 2.45% | negligible |
| RR32 | 25.13 | 62.8% | 0.06 | 1.39 | 26.58 | 66.45% | 1.45 | 3.63% | negligible |
| RR33 | 21.81 | 54.5% | 0.06 | 1.06 | 22.93 | 57.33% | 1.12 | 2.80% | negligible |
| RR34 | 27.61 | 69.0% | 0.06 | 1.66 | 29.33 | 73.33% | 1.72 | 4.30% | negligible |
| RR35 | 22.06 | 55.2% | 0.06 | 1.09 | 23.21 | 58.02% | 1.15 | 2.87% | negligible |



| Receptor | | Do-Minimum | | LSEP | | Do-Something | | | Impact |
|----------|-------|------------|----------|-------------------|-------|--------------|-------|--------|--------------------------------|
| | | | Stack PC | Road contribution | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| RR36 | 25.83 | 64.6% | 0.06 | 1.45 | 27.34 | 68.35% | 1.51 | 3.77% | negligible |
| RR37 | 22.1 | 55.3% | 0.06 | 0.35 | 22.51 | 56.27% | 0.41 | 1.02% | negligible |
| RR38 | 20.03 | 50.1% | 0.06 | 0.52 | 20.61 | 51.51% | 0.58 | 1.44% | negligible |
| RR39 | 17.56 | 43.9% | 0.05 | 0.49 | 18.10 | 45.26% | 0.54 | 1.36% | negligible |
| RR40 | 18.9 | 47.3% | 0.03 | 1.12 | 20.05 | 50.13% | 1.15 | 2.88% | negligible |
| RR41 | 17.91 | 44.8% | 0.03 | 0.97 | 18.91 | 47.28% | 1.00 | 2.50% | negligible |
| RR42 | 16.41 | 41.0% | 0.03 | 0.74 | 17.18 | 42.94% | 0.77 | 1.92% | negligible |
| RR43 | 17.02 | 42.6% | 0.02 | 0.86 | 17.90 | 44.76% | 0.88 | 2.21% | negligible |
| RR44 | 15.14 | 37.9% | 0.02 | 0.55 | 15.71 | 39.27% | 0.57 | 1.42% | negligible |
| RR45 | 15.24 | 38.1% | 0.02 | 0.58 | 15.84 | 39.59% | 0.60 | 1.49% | negligible |
| RR46 | 16.35 | 40.9% | 0.20 | 0.03 | 16.58 | 41.44% | 0.23 | 0.57% | negligible |
| RR47 | 15.96 | 39.9% | 0.20 | 0.03 | 16.19 | 40.48% | 0.23 | 0.58% | negligible |
| RR48 | 16.33 | 40.8% | 0.21 | 0.03 | 16.57 | 41.42% | 0.24 | 0.59% | negligible |
| R3 | 15.31 | 38.28% | 0.20 | 0.03 | 15.54 | 38.84% | 0.23 | 0.56% | negligible |
| R8 | 20.47 | 51.18% | 0.37 | 0.13 | 20.97 | 52.42% | 0.50 | 1.24% | negligible |
| R9 | 30.35 | 75.88% | 0.34 | 0.22 | 30.91 | 77.28% | 0.56 | 1.41% | negligible |
| R10 | 17.56 | 43.90% | 0.19 | 0.10 | 17.85 | 44.63% | 0.29 | 0.73% | negligible |
| R11 | 14.99 | 37.48% | 0.07 | 0.25 | 15.31 | 38.26% | 0.32 | 0.79% | negligible |
| R12 | 14.86 | 37.15% | 0.02 | 0.51 | 15.39 | 38.46% | 0.53 | 1.31% | negligible |



| Receptor | | Do-Minimum | | LSEP | | Do-Something | | | Impact |
|----------|-------|------------|----------------------------|-------|-------|---------------------|-------|--------|--------------------------------|
| | | | Stack PC Road contribution | | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| R13 | 12.94 | 32.35% | 0.04 | 0.11 | 13.09 | 32.72% | 0.15 | 0.37% | negligible* |
| R20 | 18.11 | 45.28% | 0.09 | 0.15 | 18.35 | 45.87% | 0.24 | 0.60% | negligible |
| R25 | 12.05 | 30.13% | 0.11 | 0.02 | 12.18 | 30.46% | 0.13 | 0.33% | negligible* |
| R27 | 16.60 | 41.50% | 0.29 | 0.08 | 16.97 | 42.42% | 0.37 | 0.92% | negligible |

Note:

Assumes 70% conversion of NOx to NO2 for process emissions

Road traffic NOx converted to NO2 using the LAQM calculator

^{*}Negligible irrespective of the total concentration.



Table 10: Annual Mean Particulate Matter as PM_{10} – Worst Case

| Receptor | I | Do-Minimum | | LSEP | | Do-Something | | | Impact |
|----------|-------|------------|----------|-------------------|-------|--------------|-------|--------|--------------------------------|
| | | | Stack PC | Road contribution | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| RR1 | 12.44 | 31.1% | 0.01 | 0.02 | 12.47 | 31.18% | 0.03 | 0.08% | negligible* |
| RR2 | 12.40 | 31.0% | 0.01 | 0.02 | 12.43 | 31.08% | 0.03 | 0.07% | negligible* |
| RR3 | 12.40 | 31.0% | 0.01 | 0.02 | 12.43 | 31.07% | 0.03 | 0.07% | negligible* |
| RR4 | 12.39 | 31.0% | 0.01 | 0.02 | 12.42 | 31.06% | 0.03 | 0.07% | negligible* |
| RR5 | 12.79 | 32.0% | 0.01 | 0.02 | 12.83 | 32.07% | 0.04 | 0.09% | negligible* |
| RR6 | 11.70 | 29.2% | 0.01 | 0.01 | 11.71 | 29.29% | 0.02 | 0.05% | negligible* |
| RR7 | 13.66 | 34.2% | 0.01 | 0.04 | 13.72 | 34.29% | 0.05 | 0.13% | negligible* |
| RR8 | 13.68 | 34.2% | 0.01 | 0.04 | 13.74 | 34.35% | 0.05 | 0.14% | negligible* |
| RR9 | 13.17 | 32.9% | 0.01 | 0.03 | 13.21 | 33.03% | 0.05 | 0.12% | negligible* |
| RR10 | 13.28 | 33.2% | 0.01 | 0.03 | 13.32 | 33.31% | 0.05 | 0.12% | negligible* |
| RR11 | 13.34 | 33.3% | 0.01 | 0.03 | 13.39 | 33.47% | 0.05 | 0.12% | negligible* |
| RR12 | 12.39 | 31.0% | 0.01 | 0.02 | 12.42 | 31.06% | 0.03 | 0.07% | negligible* |
| RR13 | 12.44 | 31.1% | 0.01 | 0.02 | 12.46 | 31.16% | 0.03 | 0.07% | negligible* |
| RR14 | 11.41 | 28.5% | <0.01 | 0.01 | 11.42 | 28.56% | 0.01 | 0.03% | negligible* |
| RR15 | 13.21 | 33.0% | <0.01 | 0.03 | 13.25 | 33.12% | 0.04 | 0.09% | negligible* |
| RR16 | 13.37 | 33.4% | <0.01 | 0.04 | 13.41 | 33.51% | 0.04 | 0.10% | negligible* |
| RR17 | 13.49 | 33.7% | <0.01 | 0.04 | 13.53 | 33.83% | 0.04 | 0.11% | negligible* |
| RR18 | 13.55 | 33.9% | <0.01 | 0.04 | 13.60 | 33.99% | 0.04 | 0.11% | negligible* |



| Receptor | | Do-Minimum | | LSEP | | Do-Something | | | Impact |
|----------|-------|------------|----------|-------------------|-------|---------------------|-------|--------|--------------------------------|
| | | | Stack PC | Road contribution | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| RR19 | 13.34 | 33.4% | <0.01 | 0.05 | 13.39 | 33.48% | 0.05 | 0.13% | negligible* |
| RR20 | 13.12 | 32.8% | <0.01 | 0.05 | 13.18 | 32.94% | 0.05 | 0.13% | negligible* |
| RR21 | 12.86 | 32.2% | <0.01 | 0.04 | 12.91 | 32.26% | 0.04 | 0.10% | negligible* |
| RR22 | 12.83 | 32.1% | <0.01 | 0.04 | 12.87 | 32.19% | 0.04 | 0.11% | negligible* |
| RR23 | 13.08 | 32.7% | <0.01 | 0.05 | 13.14 | 32.84% | 0.05 | 0.14% | negligible* |
| RR24 | 13.53 | 33.8% | <0.01 | 0.12 | 13.65 | 34.13% | 0.12 | 0.30% | negligible* |
| RR25 | 14.00 | 35.0% | <0.01 | 0.07 | 14.08 | 35.19% | 0.08 | 0.19% | negligible* |
| RR26 | 13.40 | 33.5% | <0.01 | 0.13 | 13.53 | 33.83% | 0.13 | 0.33% | negligible* |
| RR27 | 12.82 | 32.1% | <0.01 | 0.10 | 12.92 | 32.30% | 0.10 | 0.25% | negligible* |
| RR28 | 14.43 | 36.1% | <0.01 | 0.28 | 14.71 | 36.77% | 0.28 | 0.70% | negligible |
| RR29 | 13.15 | 32.9% | <0.01 | 0.16 | 13.31 | 33.27% | 0.16 | 0.40% | negligible* |
| RR30 | 14.24 | 35.6% | <0.01 | 0.26 | 14.51 | 36.26% | 0.27 | 0.67% | negligible |
| RR31 | 12.72 | 31.8% | <0.01 | 0.13 | 12.86 | 32.14% | 0.13 | 0.33% | negligible* |
| RR32 | 13.26 | 33.1% | <0.01 | 0.20 | 13.46 | 33.66% | 0.20 | 0.51% | negligible |
| RR33 | 12.68 | 31.7% | <0.01 | 0.15 | 12.84 | 32.09% | 0.15 | 0.38% | negligible* |
| RR34 | 13.72 | 34.3% | <0.01 | 0.25 | 13.96 | 34.91% | 0.25 | 0.62% | negligible |
| RR35 | 12.71 | 31.8% | <0.01 | 0.15 | 12.87 | 32.18% | 0.16 | 0.39% | negligible* |
| RR36 | 13.38 | 33.5% | <0.01 | 0.21 | 13.60 | 33.99% | 0.22 | 0.54% | negligible |
| RR37 | 12.56 | 31.4% | <0.01 | 0.06 | 12.62 | 31.54% | 0.06 | 0.15% | negligible* |



| Receptor | | Do-Minimum | | LSEP | | Do-Something | | | Impact |
|----------|-------|------------|----------|-------------------|-------|---------------------|-------|--------|--------------------------------|
| | | | Stack PC | Road contribution | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| RR38 | 12.28 | 30.7% | <0.01 | 0.09 | 12.37 | 30.92% | 0.09 | 0.22% | negligible* |
| RR39 | 11.91 | 29.8% | <0.01 | 0.08 | 11.99 | 29.98% | 0.08 | 0.21% | negligible* |
| RR40 | 12.24 | 30.6% | <0.01 | 0.20 | 12.44 | 31.10% | 0.20 | 0.50% | negligible* |
| RR41 | 12.06 | 30.2% | <0.01 | 0.17 | 12.23 | 30.58% | 0.17 | 0.43% | negligible* |
| RR42 | 11.79 | 29.5% | <0.01 | 0.13 | 11.92 | 29.80% | 0.13 | 0.32% | negligible* |
| RR43 | 11.91 | 29.8% | <0.01 | 0.15 | 12.06 | 30.14% | 0.15 | 0.37% | negligible* |
| RR44 | 11.57 | 28.9% | <0.01 | 0.09 | 11.67 | 29.16% | 0.09 | 0.23% | negligible* |
| RR45 | 11.59 | 29.0% | <0.01 | 0.10 | 11.69 | 29.22% | 0.10 | 0.24% | negligible* |
| RR46 | 11.51 | 28.8% | 0.01 | <0.01 | 11.52 | 28.80% | 0.01 | 0.03% | negligible* |
| RR47 | 11.40 | 28.5% | 0.01 | <0.01 | 11.42 | 28.54% | 0.01 | 0.03% | negligible* |
| RR48 | 11.39 | 28.5% | 0.01 | <0.01 | 11.40 | 28.49% | 0.01 | 0.03% | negligible* |
| R3 | 11.32 | 28.29% | 0.01 | 0.00 | 11.33 | 28.32% | 0.01 | 0.03% | negligible* |
| R8 | 12.72 | 31.80% | 0.01 | 0.02 | 12.76 | 31.89% | 0.04 | 0.10% | negligible* |
| R9 | 14.34 | 35.84% | 0.01 | 0.05 | 14.40 | 36.00% | 0.06 | 0.16% | negligible* |
| R10 | 12.31 | 30.77% | 0.01 | 0.02 | 12.34 | 30.84% | 0.03 | 0.07% | negligible* |
| R11 | 11.49 | 28.73% | <0.01 | 0.04 | 11.53 | 28.82% | 0.04 | 0.10% | negligible* |
| R12 | 11.52 | 28.81% | <0.01 | 0.09 | 11.61 | 29.02% | 0.09 | 0.21% | negligible* |
| R13 | 11.17 | 27.91% | <0.01 | 0.02 | 11.18 | 27.96% | 0.02 | 0.04% | negligible* |
| R20 | 11.94 | 29.86% | <0.01 | 0.02 | 11.97 | 29.91% | 0.02 | 0.06% | negligible* |



| Receptor | r Do-Minimum | | LSEP | | Do-Something | | Impact | | |
|----------|--------------|--------|----------|-------------------|--------------|--------|--------|--------|--------------------------------|
| | | | Stack PC | Road contribution | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| R25 | 11.39 | 28.48% | <0.01 | <0.01 | 11.40 | 28.50% | 0.01 | 0.02% | negligible* |
| R27 | 12.05 | 30.12% | 0.01 | 0.02 | 12.07 | 30.19% | 0.03 | 0.07% | negligible* |

Note:

Impacts calculated as a percentage of the AQAL of 40 $\mu g/m^3$

Table 11: Annual Mean Particulate Matter as PM_{2.5} – Worst Case

| Receptor | | Do-Minimum | | LSEP | | Do-Something | | | Impact |
|----------|-------|------------|----------|-------------------|-------|---------------------|-------|--------|--------------------------------|
| | | | Stack PC | Road contribution | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| RR1 | 7.91 | 39.56% | 0.01 | 0.01 | 7.94 | 39.68% | 0.02 | 0.12% | negligible* |
| RR2 | 7.91 | 39.53% | 0.01 | 0.01 | 7.93 | 39.64% | 0.02 | 0.11% | negligible* |
| RR3 | 7.90 | 39.51% | 0.01 | 0.01 | 7.92 | 39.62% | 0.02 | 0.11% | negligible* |
| RR4 | 7.90 | 39.49% | 0.01 | 0.01 | 7.92 | 39.60% | 0.02 | 0.11% | negligible* |
| RR5 | 8.14 | 40.70% | 0.01 | 0.01 | 8.17 | 40.83% | 0.03 | 0.13% | negligible* |
| RR6 | 7.48 | 37.39% | 0.01 | 0.00 | 7.50 | 37.48% | 0.02 | 0.08% | negligible* |
| RR7 | 8.59 | 42.93% | 0.01 | 0.02 | 8.62 | 43.12% | 0.04 | 0.19% | negligible* |
| RR8 | 8.60 | 43.01% | 0.01 | 0.02 | 8.64 | 21.60% | 0.04 | 0.19% | negligible* |

^{*}Negligible irrespective of the total concentration.



| Receptor | | Do-Minimum | | LSEP | | Do-Something | | | Impact |
|----------|-------|------------|----------|-------------------|-------|--------------|-------|--------|--------------------------------|
| | | | Stack PC | Road contribution | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor |
| RR9 | 8.29 | 41.44% | 0.01 | 0.02 | 8.32 | 20.80% | 0.03 | 0.17% | negligible* |
| RR10 | 8.35 | 41.76% | 0.01 | 0.02 | 8.39 | 20.96% | 0.03 | 0.17% | negligible* |
| RR11 | 8.39 | 41.95% | 0.01 | 0.02 | 8.42 | 21.06% | 0.03 | 0.17% | negligible* |
| RR12 | 7.82 | 39.10% | 0.01 | 0.01 | 7.84 | 19.60% | 0.02 | 0.10% | negligible* |
| RR13 | 7.85 | 39.23% | 0.01 | 0.01 | 7.86 | 19.66% | 0.02 | 0.10% | negligible* |
| RR14 | 7.41 | 37.06% | <0.01 | 0.01 | 7.42 | 18.55% | 0.01 | 0.05% | negligible* |
| RR15 | 8.57 | 42.85% | <0.01 | 0.02 | 8.59 | 21.49% | 0.02 | 0.12% | negligible* |
| RR16 | 8.67 | 43.36% | <0.01 | 0.02 | 8.70 | 21.75% | 0.03 | 0.13% | negligible* |
| RR17 | 8.75 | 43.75% | <0.01 | 0.02 | 8.78 | 21.95% | 0.03 | 0.14% | negligible* |
| RR18 | 8.79 | 43.95% | <0.01 | 0.02 | 8.82 | 22.05% | 0.03 | 0.15% | negligible* |
| RR19 | 8.55 | 42.77% | <0.01 | 0.03 | 8.59 | 21.46% | 0.03 | 0.16% | negligible* |
| RR20 | 8.43 | 42.13% | <0.01 | 0.03 | 8.46 | 21.14% | 0.03 | 0.16% | negligible* |
| RR21 | 8.29 | 41.43% | <0.01 | 0.02 | 8.31 | 20.78% | 0.03 | 0.13% | negligible* |
| RR22 | 8.26 | 41.32% | <0.01 | 0.02 | 8.29 | 20.73% | 0.03 | 0.14% | negligible* |
| RR23 | 8.41 | 42.07% | <0.01 | 0.03 | 8.45 | 21.12% | 0.03 | 0.17% | negligible* |
| RR24 | 8.68 | 43.38% | <0.01 | 0.07 | 8.75 | 21.87% | 0.07 | 0.36% | negligible* |
| RR25 | 8.95 | 44.74% | <0.01 | 0.04 | 9.00 | 22.49% | 0.05 | 0.23% | negligible* |
| RR26 | 8.60 | 42.98% | <0.01 | 0.08 | 8.67 | 21.69% | 0.08 | 0.40% | negligible* |
| RR27 | 8.25 | 41.23% | <0.01 | 0.06 | 8.31 | 20.77% | 0.06 | 0.30% | negligible* |



| Receptor | Do-Minimum | | LSEP | | | Do-Something | | Impact | | |
|----------|------------|--------|----------|-------------------|-------|--------------|-------|--------|--------------------------------|--|
| | | | Stack PC | Road contribution | | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor | |
| RR28 | 9.19 | 45.96% | <0.01 | 0.17 | 9.36 | 23.40% | 0.17 | 0.84% | negligible | |
| RR29 | 8.44 | 42.18% | <0.01 | 0.10 | 8.53 | 21.33% | 0.10 | 0.49% | negligible* | |
| RR30 | 9.08 | 45.39% | <0.01 | 0.16 | 9.24 | 23.10% | 0.16 | 0.81% | negligible | |
| RR31 | 8.19 | 40.93% | <0.01 | 0.08 | 8.27 | 20.67% | 0.08 | 0.41% | negligible* | |
| RR32 | 8.59 | 42.95% | <0.01 | 0.12 | 8.71 | 21.78% | 0.12 | 0.62% | negligible | |
| RR33 | 8.25 | 41.24% | <0.01 | 0.09 | 8.34 | 20.85% | 0.09 | 0.46% | negligible* | |
| RR34 | 8.86 | 44.29% | <0.01 | 0.15 | 9.01 | 22.52% | 0.15 | 0.75% | negligible | |
| RR35 | 8.27 | 41.34% | <0.01 | 0.09 | 8.36 | 20.91% | 0.10 | 0.48% | negligible* | |
| RR36 | 8.66 | 43.31% | <0.01 | 0.13 | 8.79 | 21.98% | 0.13 | 0.66% | negligible | |
| RR37 | 8.19 | 40.95% | <0.01 | 0.03 | 8.23 | 20.57% | 0.04 | 0.18% | negligible* | |
| RR38 | 8.02 | 40.10% | <0.01 | 0.05 | 8.07 | 20.18% | 0.05 | 0.27% | negligible* | |
| RR39 | 7.80 | 38.98% | <0.01 | 0.05 | 7.85 | 19.62% | 0.05 | 0.25% | negligible* | |
| RR40 | 7.98 | 39.92% | <0.01 | 0.12 | 8.10 | 20.26% | 0.12 | 0.59% | negligible | |
| RR41 | 7.88 | 39.40% | <0.01 | 0.10 | 7.98 | 19.95% | 0.10 | 0.51% | negligible | |
| RR42 | 7.72 | 38.61% | <0.01 | 0.07 | 7.80 | 19.49% | 0.08 | 0.38% | negligible* | |
| RR43 | 7.79 | 38.94% | <0.01 | 0.09 | 7.88 | 19.69% | 0.09 | 0.44% | negligible* | |
| RR44 | 7.59 | 37.96% | <0.01 | 0.05 | 7.65 | 19.12% | 0.06 | 0.28% | negligible* | |
| RR45 | 7.60 | 38.01% | <0.01 | 0.06 | 7.66 | 19.15% | 0.06 | 0.29% | negligible* | |
| RR46 | 7.55 | 37.76% | 0.01 | <0.01 | 7.56 | 18.91% | 0.01 | 0.05% | negligible* | |



| Receptor | Do-Minimum S | | | LSEP | | Do-Something | | Impact | | |
|----------|--------------|--------|----------|-------------------|-------|--------------|-------|--------|--------------------------------|--|
| | | | Stack PC | Road contribution | | | | | | |
| | μg/m³ | % AQAL | μg/m³ | μg/m³ | μg/m³ | % AQAL | μg/m³ | % AQAL | Magnitude of change descriptor | |
| RR47 | 7.49 | 37.46% | 0.01 | <0.01 | 7.50 | 18.76% | 0.01 | 0.05% | negligible* | |
| RR48 | 7.49 | 37.44% | 0.01 | <0.01 | 7.50 | 18.75% | 0.01 | 0.06% | negligible* | |
| R3 | 7.44 | 37.19% | 0.01 | <0.01 | 7.45 | 37.25% | 0.01 | 0.05% | negligible* | |
| R8 | 8.02 | 40.09% | 0.01 | 0.01 | 8.05 | 40.23% | 0.03 | 0.15% | negligible* | |
| R9 | 8.99 | 44.96% | 0.01 | 0.03 | 9.03 | 45.17% | 0.04 | 0.21% | negligible* | |
| R10 | 7.77 | 38.84% | 0.01 | 0.01 | 7.79 | 38.94% | 0.02 | 0.10% | negligible* | |
| R11 | 7.55 | 37.74% | <0.01 | 0.02 | 7.57 | 37.85% | 0.02 | 0.12% | negligible* | |
| R12 | 7.56 | 37.81% | <0.01 | 0.05 | 7.61 | 19.03% | 0.05 | 0.26% | negligible* | |
| R13 | 7.35 | 36.77% | <0.01 | 0.01 | 7.37 | 18.41% | 0.01 | 0.06% | negligible* | |
| R20 | 7.73 | 38.65% | <0.01 | 0.01 | 7.75 | 19.37% | 0.02 | 0.08% | negligible* | |
| R25 | 7.38 | 36.90% | <0.01 | <0.01 | 7.39 | 18.47% | 0.01 | 0.03% | negligible* | |
| R27 | 7.68 | 38.38% | 0.01 | 0.01 | 7.70 | 19.24% | 0.02 | 0.10% | negligible* | |

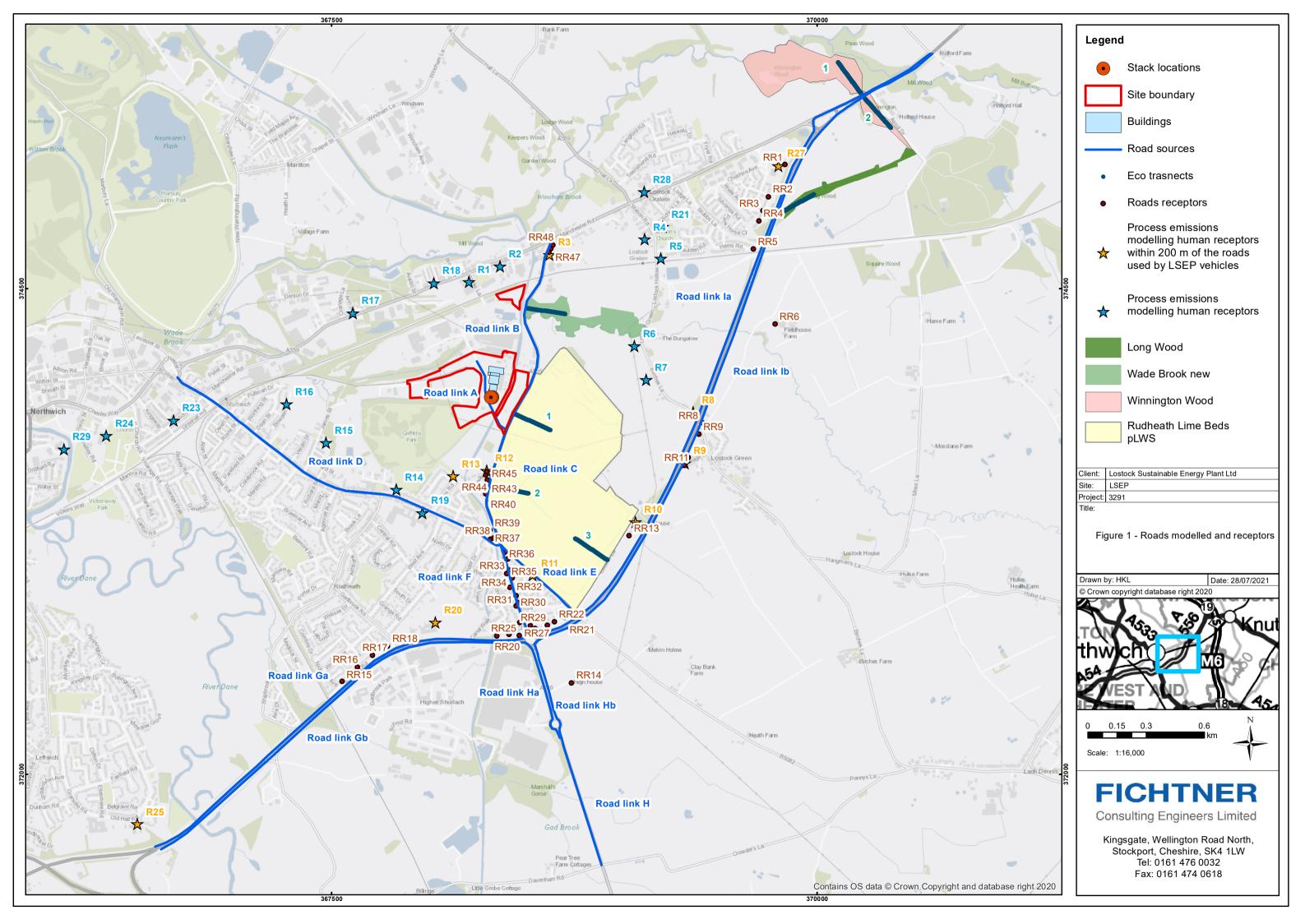
Note:

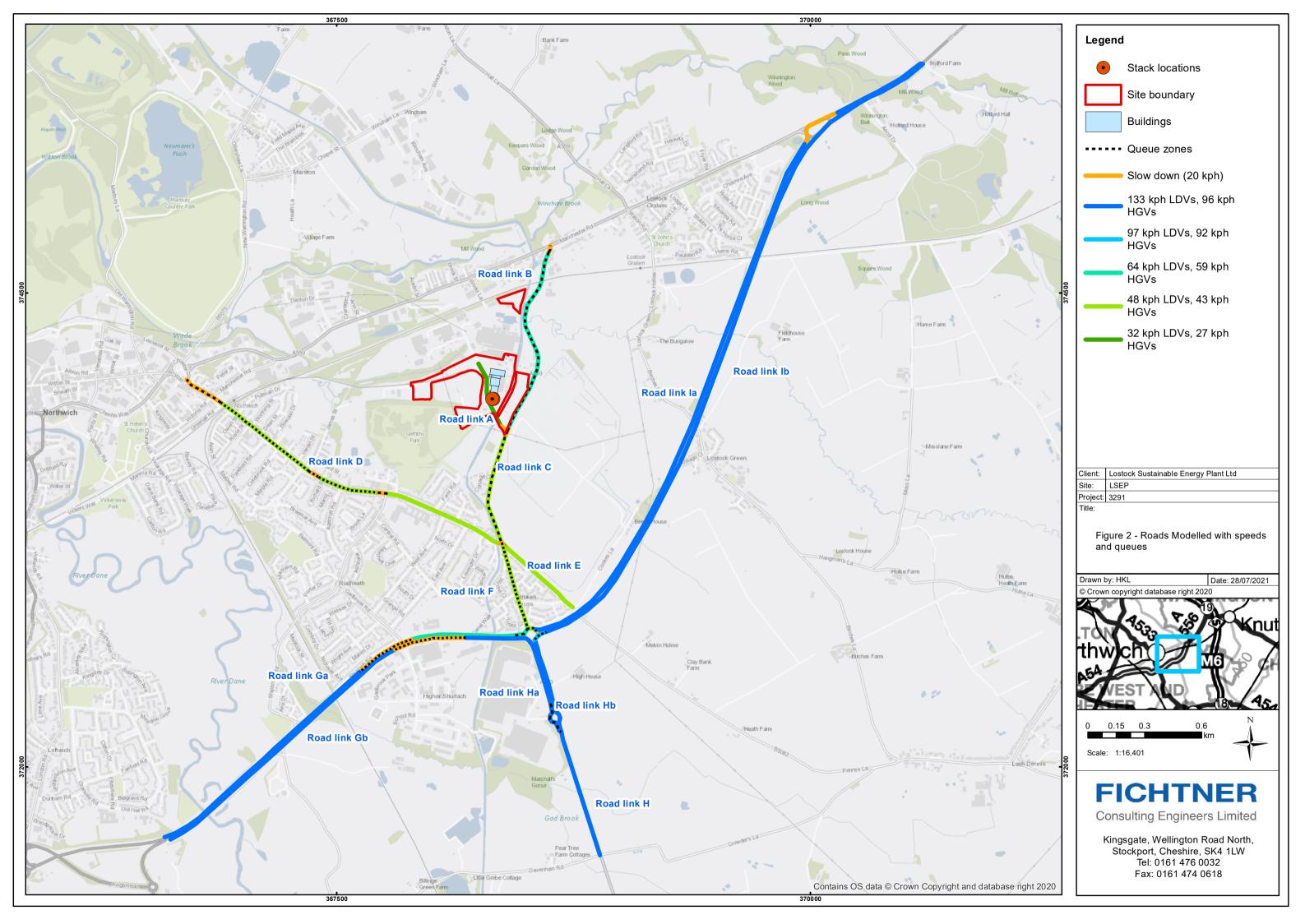
Impacts calculated as a percentage of the AQAL of 20 $\mu g/m^3$

^{*}Negligible irrespective of the total concentration.



C Figures





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